Soil chemical properties of opencast coal mining site in Indonesia and its effect on plant growth

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ABSTRACT

Opencast coal mining causes severe ecosystem disturbances, such as soil degradation, excessive waste production, overburden formation, and possibly reforestation inhibition. In this study, we assessed the effect of opencast coal mining on soil chemical properties and plant growth in soil from natural forests and post coal mining sites (a 16-year-old coal mining site in Lati and a 5-year-old coal mining site in Sambarata) in Berau Regency, East Kalimantan, Indonesia. Soil pH, total carbon, total nitrogen, available phosphate, exchangeable cations, and cation exchange capacity were measured. Plant growth and shoot phosphorus, potassium, calcium, magnesium, and iron concentrations were determined as well. Soil from the post coal mining sites had lower pH, total carbon, total nitrogen, available phosphate, potassium, sodium, calcium, magnesium, and cation exchange capacity than soil from the natural forests. Reduction of shoot nutrient content and plant biomass was observed in *Sorghum bicolor* grown in soil from both Lati and Sambarata post coal mining sites. By calculating the correlation between soil chemical properties and plant growth, it was found that all the soil chemical properties are limiting factors of plant growth in Lati. The results revealed that opencast coal mining reduced soil fertility and inhibited plant growth, which hinders successful reforestation. Thus, monitoring, evaluation, and restoration of post coal mining area is recommended to prevent more severe environmental damage.

Key words: Opencast coal mining, Soil degradation, Restoration, Growth limitation, Kalimantan

Introduction

Coal mining is one of the most important industries in Indonesia (Suryantoro and Manaf, 2002). According to US Energy Information Administration (www.eia.gov), Indonesia ranks 5th among the world's coal producers, with total production reaching approximately 414 million tons in 2011. The mining industry, including coal mining, accounts for approximately 4-5% of Indonesia's gross domestic product (GDP). Among the minerals produced in Indonesia, coal has the highest production, and its price is second only to that of gold (Directorate General of Minerals and Coal, 2012). Coal mining is widespread in the six main islands of Indonesia, namely, Sumatra, Kalimantan, Java, Sulawesi, Maluku, and Irian Jaya (Syahrial *et al.*, 2012).

Coal mining is known to contribute to environmental degradation. Opencast mining is conducted in Indonesia because coal is mostly located close to the surface and under natural forest areas (Resosudarmo et al., 2009). The process of opencast coal mining, which involves logging trees, stripping topsoil, removing overburden, and exploiting minerals (Ghose and Majee, 1998), leads to deforestation and severely damages both local and global ecosystems (Resosudarmo et al., 2009). The impact of opencast coal mining is greater than that of underground coal mining as there is no aboveground disturbance by underground mining. In addition to deforestation, opencast coal mining results in excessive waste production (Bian et al., 2010), overburden (Dowarah et al., 2009), increased soil erosion (Braghina et al., 2010), and reduction of soil essential nutrients (Sheoran et al., 2010), all of which further deteriorate soil fertility (Ghose, 2004). Finally, opencast coal mining may alter soil physical, chemical, and biological properties, leading to extensive soil degradation (Juwarkar and Jambhulkar, 2008).

Soil degradation by opencast mining decreases plant species richness, inhibits plant development (Burger and Zipper, 2002), damages aboveground flora (Monjezi *et al.*, 2009), and destroys native vegetation (Chen *et al.*, 1998). This adverse impact of soil degradation on aboveground vegetation has direct negative consequences on the habitat and biota of fauna (Monjezi *et al.*, 2009), wildlife, community, and finally, the ecosystem (Peplow and Edmonds, 2005). Opencast mining also causes heavy metal contamination of the adjacent land (Kien *et al.*, 2010).

Clearly, the restoration of opencast mining areas is a must. Natural forest recovery occurs very slowly (Jha and Singh, 1991) and it is estimated that approximately 200 years is required to reach the level of native forests (Srivastava *et al.*, 1989). Revegetation is usually adopted for soil and environment stabilization in areas degraded by opencast mining (Wong, 2003). However, most revegetation attempts are inhibited by problems concerning soil chemical or physical properties. The restoration of degraded land in opencast mining areas can be achieved by the determination of soil chemical, physical, and biological properties (Sheoran *et al.*, 2010).

The evaluation of soil chemical properties in opencast coal mining areas is important. The removal of topsoil, the erosion of exposed areas (Carrol and Tucker, 2000), low soil organic matter, and the lack of nutrients create problems for revegetation on degraded mine spoils (Makineci *et al.* 2011). Few studies have investigated the impact of opencast coal mining on soil chemical properties in Indonesia. This is a serious concern in Kalimantan because this area is the second largest coal producer in Indonesia (Syahrial *et al.*, 2012) and suffers from the highest rate of deforestation (FAO, 2009).

In this study, we focused on the impact of opencast coal mining on soil chemical properties and the potential limiting factors of reforestation in Berau Regency, East Kalimantan, Indonesia. We compared soil chemical properties and plant growth using soil from post coal mining sites and natural forests. As bio-indicator, we used Sorghum bicolor to evaluate the limiting factors of plant growth in post coal mining site, because this plant is able to grow in a short time and adapts well to wide ranges of soil pH, high Al stress, and low-fertility soil, such as ultisol soil (Flores et al., 1988; Flores et al., 1991). We investigated the relationships among opencast mining, soil chemical properties, and plant growth. In particular, we addressed the following questions: (1) Does opencast mining change soil chemical properties? (2) Does the soil chemical property in post mining site effect on plant growth?

Materials and Methods

Study site and soil sampling

The study site was located in PT. Berau Coal and lies between the longitude (01°52'26.74" -02°25'09.78" N; 117°07'44.52" - 117°38'26.46" E), Berau Regency East Kalimantan Indonesia. PT Berau Coal is the first contractor of coal mining in the country to operate a concession area in Berau approximately 121.804 ha. The climate is of tropictype with a mean annual rainfall of 3132 mm yr⁻¹. A mean daily minimum temperature is range of 22 °C and a mean daily maximum of 35 °C. A mean daily minimum velocity is range of 5.8 knot and a mean daily maximum of 11.42 knot. A mean of daily humidity is 87% and a mean of daily sunshine duration is 44%. This region is located in tropical rain forest zone with Dipterocarpacea forest as its typical vegetation. Two areas of two ages of mining were selected which encompassed two post coal mining sites and their respective natural forests (pre-mining site): a 16-year-old post coal mining site (02°15'N; 117°35'E) and its natural forest (02°15'N; 117°34'E) at Lati, and a 5-year-old post coal mining site (02°10'N; 117°24'E) and its natural forest (02°10'N; 117°24'E) at Sambarata, in Berau Regency, East Kalimantan, Indonesia. Both natural forests were dominated by Anisoptera costata (Dipterocarpaceae), whereas there was no vegetation covering the post coal mining sites. Lati mine is the biggest mine site at PT Berau Coal and laid 35 km from the East of Tanjung Redeb, a capital city of Berau Regency. Lati mine started the operation in 1993 with a production capacity of 15 Mt of coal per year and more than 120 x 10⁶ m³ of overburden movement annually. While in Sambarata mine, coal operation was started in 2001. The coal deposits are mined using open pit mining method involves the complete removal of vegetation, top soil, and underlying gravel layer. After the removal of vegetation, the top soil and underlying gravel layer or overburden were collected and used as backfilling in post coal mining site.

Soils were ultisol and collected representatively in uniform field in a simple random pattern across the field. Field had flat slope, dry, no erosion, and without fertilizer or manure application history. The top soil (0-10 cm) and under layer gravel (10-30 cm) were collected, weighed approximately 0.5 kg, and composited. Soils were collected in approximately 90 x 90 m² of wide sampling area with 30 m distance between soil sampling. Five replications of soil samples were collected at each site. In Lathi area, soil texture in natural forest was silty clay loam, and soil texture in post coal mining site was clay loam. While in Sambarata area, soil texture in natural forest and post coal mining site was loam.

Soil chemical analysis

Soil was air-dried and passed through a < 2 mm sieve. The passed dried soil was used for analysis of pH (H₂O) and pH (KCl). Available phosphate (P) (Truog 1930) was extracted with 0.001 M sulfuric acid solutions and analyzed by the ammonium molybdate method (Olsen and Sommers 1982). Total carbon (TC) and total nitrogen (TN) were determined by a C:N analyzer (Sumigraph NC-220F, Tokyo). Exchangeable potassium (K), sodium (Na), magnesium (Mg), and calcium (Ca) were extracted with 1M (pH 7) ammonium acetate solution and their concentrations were determined using an atomic absorption spectrophotometer (Hitachi model Z-5000 series Polarized Zeeman, Tokyo). After removing excess NH⁺₄, the sample was extracted with 100 g L⁻¹ KCl solution and the supernatant was used to determine cation exchange capacity (CEC) using the semi-micro Schöllenberger method. Base saturation (BS) was calculated by dividing the sum of exchangeable cations (K, Na, Mg, Ca) by CEC and multiplying the result by 100%.

Plant growth

Owing to the limiting number of soil samples and the short time of plant cultivation (3 months), a fifty ml cylindrical pot (3 cm diameter x 11 cm height) was used and enough to grow Sorghum bicolor (CV) New sorgo No.2 in one hundred grams of fresh soil sample. Five seeds of Sorghum bicolor (CV) New sorgo No.2 were sown. Seedlings were thinned to three in each pot after germination and grown under greenhouse conditions at Yamagata University, Tsuruoka, Japan (38°44'N, 139°50'E) for 90 days. Five replications were made. The seedlings in the 50 ml cylinder pots were placed randomly in the greenhouse and approximately 15 mL of tap water was applied once every two days. No fertilizer was applied. In three months after sowing, shoots and roots were harvested, washed with tap water and deionized water, separated, and oven-dried at 70 °C for 72 hours before weighing. The dried shoots were ground and digested with a solution of nitric acid (HNO₃), perchloric acid (HClO₄), and sulfuric acid (H_2SO_4) (5 : 2 : 1 volume). P concentration in the digestion solution was determined colorimetrically with the vanadomolybdate-yellow assay (Olsen and Sommers, 1982) by employing a spectrophotometer (Hitachi U-2900, Tokyo) with the absorbance set at 880 nm. Shoot K, Mg, Fe, and Ca concentrations in the digestion solution were determined with atomic absorption spectrophotometer (Hitachi 170-50, Tokyo) using hollow cathode lamps. Shoot P, K, Mg, Fe, and Ca contents were calculated by multiplying shoot P, K, Mg, Fe, and Ca concentrations by shoot dry weights.

Statistical analysis

Statistical significance was analyzed using Kaleida Graph 4.1 software (Synergy Software 2012, USA) and means of groups were compared using the student t-test (P < 0.05).

Results

Chemical properties of soil from natural forests and post coal mining sites

Marked degradation of the chemical properties of soil from Lathi and Sambarata post coal mining sites

was noted (Table 1). TC concentration and CEC were lower in soil from the post coal mining sites than in soil from the natural forests in both Lati and Sambarata. The reduction of soil TC concentration was approximately 89% in Sambarata and 92% in Lati. CEC was decreased by approximately 31% and 37% in soil from Lati and Sambarata post coal mining sites, respectively (Table 3).

Large reductions of TN, available P, and exchangeable K, Mg, Ca, and Na concentrations were observed in soil from the post coal mining sites in both Lathi and Sambarata (Table 1). TN concentration was decreased by approximately 49% and 65% in soil from the post coal mining sites in Sambarata and Lati, respectively (Table 3). Available P was decreased to approximately 67% and 57% in soil from the post coal mining sites in Lati and Sambarata, respectively (Table 3). Soil K, Ca, and Mg concentrations were decreased by approximately 61%, 95%, and 91%, respectively, in Lati.

The pH (H_2O) and pH (KCl) were decreased in soil from the post coal mining sites in both Lati and Sambarata (Table 1). The high acidity is due to the small amount of exchangeable cations. Mining decreased soil pH (H_2O) by 10% in both Lati and Sambarata, whereas it decreased soil pH (KCl) by approximately 23% in Lati and 11% in Sambarata.

The relatively large decline of TC, available P, and exchangeable K, Ca, Mg, and Na concentrations was mostly recorded in soil from the 16-year-old mining site in Lati than in soil from the 5-year-old mining site in Sambarata.

Plant growth, shoot nutrient concentration, and shoot nutrient content

To gain a better understanding of mining effects on soils and plants, growth and biomass response of *S. bicolor* were monitored and the results of which are expected to reveal the limiting factors of plant growth. Shoot dry weight of *S. bicolor* grown in soil from the post coal mining sites was lower than that grown in soil from the natural forests in both Lati and Sambarata (Fig. 1).

Shoot nutrient concentrations of *S. bicolor* grown in soil from the post coal mining sites and the natural forests in Lati and Sambarata are shown in Table 2. Use of soil from Lati post coal mining site resulted in the significant reduction of shoot nutrient contents: P was reduced by 41%; K, by 73%; Ca, by 72%; Mg, by 59%; and Fe, by 54%. Meanwhile, use of soil from Sambarata post coal mining site resulted in the Eco. Env. & Cons. 26 (August Suppl. Issue) : 2020

						Exchangeable	e cation					
Area	H,O,H	H KCI	Total Carbon	Total Nitrogen	C:N ratio	Avail. P (mg P,O ₅ kg ⁻¹)	X	Na (cmol	Ca kg^- ¹)	Mg	CEC (cmol_kg ⁻¹)	BS (%)
	1			(%))) 4))	
Lati												
Natural forest	5.08 a	5.04 a	2.23 a	0.20 a	11.09 a	10.10 a	0.64 a	0.10 a	2.80 a	2.30 a	11.60 a	76.67 a
	(0.07)	(0.08)	(0.36)	(0.008)	(1.94)	(0.08)	(0.16)	(0.005)	(0.91)	(0.34)	(1.01)	(4.58)
Post coal mining	4.60 b	3.88 b	0.19 b	0.07 b	2.72 b	$3.30 \mathrm{b}$	0.25 b	0.07 b	0.14 b	0.22 b	7.30 b	9.70 b
)	(0.11)	(0.03)	(0.02)	(0.005)	(0.33)	(0.06)	(0.02)	(0.004)	(0.01)	(0.0)	(0.59)	(2.34)
Sambarata												
Natural forest	4.61 a	3.83 a	2.84 a	0.27 a	10.44 a	9.60 a	0.41 a	0.09 a	0.32 a	0.95 a	12.88a	38.37 a
	(0.11)	(0.03)	(0.27)	(0.006)	(0.92)	(0.20)	(0.04)	(0.006)	(0.05)	(0.19)	(0.95)	(5.66)
Post coal mining	4.16 b	3.44 b	0.31 b	0.14 b	2.25 b	4.10 b	0.21 b	0.07 b	0.09 b	$0.10 \mathrm{b}$	8.88 b	$5.64 \mathrm{b}$
)	(0.08)	(0.02)	(0.03)	(0.005)	(0.18)	(0.13)	(0.02)	(0.007)	(0.01)	(0.06)	(0.78)	(1.44)
H ₂ O, water; KCl, p	otassium cł	nloride; C:N	J, carbon-to-	-nitrogen rat	io; Avail.	P, available ph	osphate; F	<, potassium	ı; Na, sodiu	m; Ca, calc	cium; Mg, mag	;nesium;
CEC, cation exchar	nge capacity	r; BS, base sí	aturation. Va	alues in pare	ntheses a	re means of five	replicates	s standard ei	ror (SE). Di	fferent lette	ers in the same	column
within the same ar	ea indicate	a statistical	ly significar	nt difference	between	post coal minin	g site and	natural fore	est (P<0.05)	according	to student t-te	st (n=5).

post coal mining sites in Lati and Sambarata, East Kalimantan, Indonesia

Table 1. Chemical properties of soil from natural forests and



Fig. 1. Shoot dry weight of *Sorghum bicolor* grown in soil from 5-year-old (Sambarata) and 16-year-old (Lati) post coal mining sites in East Kalimantan, Indonesia. * 5% level of significance; *** 0.1% level of significance by the Tukey HSD test (n=5). *Vertical bars* are standard errors of means (SE)

significant reduction of only Mg content by 75%.

Shoot K concentration was decreased by 34% in *S. bicolor* grown in soil from the post coal mining site in Lati (Table 2). Similarly, shoot Mg concentration was decreased by 63% in *S. bicolor* grown in soil from the post coal mining site in Sambarata.

Discussion

Impact of coal mining on soil chemical properties

Coal mining significantly decreased the soil chemical properties. The decrease was extensive due to the excavation process adopted by opencast coal mining, which involved the removal of cover trees, topsoil, and litter. The reduction of soil TC concentration in Sambarata and in Lati is much higher about 25% - 40% than the reduction of TC in India (Singh et al., 2012) and Ohio, USA (Shrestha and Lal, 2011), respectively. The decrease in TC concentration in soil from the post coal mining site also lowered CEC. Furthermore, mixing of the lower soil horizon during opencast mining might have lowered CEC. Those values are within the range observed in a previous study that reported a CEC reduction of approximately 26-37% in soil from a coal mining site in India (Sadhu et al., 2012).

Large depletion of TN, available P, and ex-

Area		Shoot 1	nutrient concen	itration			Shoc	t nutrient cont	ent	
	Þ.	К	Ca	Mg	Fe	Р	K	Са	Mg	Fe
	$(mg g^{-1})$		$(mg pot^1)$))	
Lati:										
Natural forest	0.38 b	14.21 a	4.61 a	7.48 a	0.15 a	0.17 a	6.25 a	2.10 a	4.60 a	0.063 a
	(0.03)	(1.09)	(06.0)	(1.97)	(0.02)	(0.01)	(0.58)	(0.45)	(0.44)	(0.006)
Post coal mining	0.58 a	9.48 b	3.58 a	4.50 a	0.18 a	0.10 b	1.66 b	0.58 b	1.90 b	0.029 b
)	(0.05)	(1.15)	(0.96)	(1.05)	(0.05)	(0.01)	(0.17)	(0.03)	(0.31)	(0.001)
Sambarata:										
Natural forest	0.32 b	8.37 a	2.48 a	10.24 a	0.49 a	0.09 a	2.28 a	0.72 a	3.02 a	0.133 a
	(0.03)	(1.33)	(0.49)	(0.69)	(0.21)	(0.01)	(0.32)	(0.16)	(0.57)	(0.054)
Post coal mining	0.85 a	8.06 a	4.28 a	$3.76 \mathrm{b}$	0.22 a	0.12 a	1.30 a	0.60 a	0.75 b	0.028 a
)	(0.13)	(0.28)	(0.76)	(1.12)	(0.05)	(0.02)	(0.32)	(0.11)	(0.33)	(0.004)
P, phosphorus; K, p in the same column	otassium; Ca, within the sa	calcium; Mg, me area indic	, magnesium; F ate a statistical	e, iron. Value ly significant	s in parenthe difference be	ses are means tween natural	of five replication of forest and po	ites standard e st coal mining	rror (SE). Diff site (P<0.05) i	erent letters according to
סומתבזוו-ו יכסי לדו-הי.										

Shoot nutrient concentrations and contents of Sorghum bicolor grown in soil from natural forests and post coal mining sites in Lati and Sambarata

Table 2.

changeable K, Mg, Ca, and Na concentrations in both Lathi and Sambarata area is due to the removal of topsoil, vegetation, and litter that plays a prominent role in supplying organic matter and nutrients to soil during opencast coal mining. As it is reported that litter from a tropical forest in Kalimantan contributes approximately 0.4-1% N (Vernimmen et al. 2007). However, the reduction of TN is smaller about 10% - 20% in comparison to N reduction in soil from a post coal mining site in India (Singh et al. 2012), and in Ohio, USA (Shrestha and Lal, 2011). This may be explained by the low denitrification rate (Vernimmen et al. (2007), low nitrification from 0 to 35% (Ohta and Effendi, 1992) and low mineralization of N in Kalimantan (Vernimmen et al., 2007) in the tropical forests of Kalimantan. Furthermore, the high rainfall (3132 mm yr⁻¹) in Tanjung Redeb (Station of Meteorology of Tanjung Redeb, Berau Regency 2008), may explain the low N (Santiago et al., 2005; Alvarez-Clare and Mack, 2011). The low C:N ratio in soil from the post coal mining site is due to mineralization of soil organic matter that results from the loss of C and N due to mining.

The impact of opencast coal mining on the reduction of available P in Lathi and Sambarata area is much higher about 30% - 40% than reduction of available P in opencast coal mining in India (Ghose 2004). The low available P in this study can be explained by the high soil acidity (pH< 5), which greatly reduces availability of P (Hazelton and Murphy, 2007). Together, these findings indicate that the high soil acidity and the low soil organic matter are the limiting factors of available P.

The reduction of soil chemical properties was also found for K, Ca, and Mg in both post coal mining in Lathi and Sambarata area. Those reductions are much higher in this study than the K reduction of 28-46% (Ghose, 2004), Ca reduction of 7-46%, and Mg reduction of 9-43% in opencast coal mining in India (Sadhu *et al.*, 2012).

Coal mining activity increased the soil acidity in Lathi and Sambarata area. The high acidity is due to

Soil chemical property	Place	Ratio (%) post mining: pre mining	Mining type	Reference
С	Lati	9	Coal	Present study
	Sambarata	11	Coal	Present study
	Ohio	14-44	Coal	Shrestha and Lal (2011)
	India	33	Coal	Singh <i>et al.</i> (2012)
CEC	Lati	69	Coal	Present study
	Sambarata	63	Coal	Present study
	India	63-74	Coal	Sadhu <i>et al.</i> (2012)
Ν	Lati	35	Coal	Present study
	Sambarata	51	Coal	Present study
	Ohio	20-47	Coal	Shrestha and Lal (2011)
	India	47	Coal	Singh <i>et al.</i> (2012)
Р	Lati	33	Coal	Present study
	Sambarata	43	Coal	Present study
	India	65-77	Coal	Ghose (2004)
Κ	Lati	39	Coal	Present study
	Sambarata	51	Coal	Present study
	India	54-72	Coal	Ghose (2004)
Ca	Lati	5	Coal	Present study
	Sambarata	28	Coal	Present study
	India	54-93	Coal	Sadhu <i>et al.</i> (2012)
Mg	Lati	9	Coal	Present study
<u> </u>	Sambarata	10	Coal	Present study
	India	57-91	Coal	Sadhu <i>et al</i> . (2012)

Table 3. Changes of soil chemical properties due to mining activities in countries having tropical climate and temperate climate, in comparison to soil from post coal mining sites in Lati and Sambarata, Indonesia

C, carbon; CEC, cation exchange capacity; N, nitrogen; P, phosphate; K, potassium; Ca, calcium; Mg, magnesium; Co, copper; Pb, lead; Zn, zinc; As, arsenic; Mn, manganese.

the small amount of exchangeable cations. The changes of soil acidity in this study are much higher than the pH reduction of approximately 2% to 4% in a dry deciduous forest in India (Sadhu *et al.*, 2012).

Generally, mining adversely affects soil chemical properties, reducing them to extremely low levels that possibly inhibit plant growth. Comparison with other opencast mining sites in countries with different climates revealed the severity of soil degradation due to opencast mining in East Kalimantan, Indonesia (Table 3). The high rainfall in Kalimantan possibly causes the high runoff that lead to the marked degradation of soil chemical properties. Rainfall in Tanjung Redeb, Berau Regency East Kalimantan measures approximately 3132 mm yr⁻¹ (Station of Meteorology of Tanjung Redeb, Berau Regency, 2008), which is much higher than rainfall in dry tropical India, which is approximately 1240-1500 mm y⁻¹ (Sadhu *et al.*, 2012). Santiago *et al.* (2005) reported that soil P, K, Ca, and Mg concentrations decreased with increasing precipitation gradients of 1800, 2300, 3100, and 3500 mm yr-1 in a lowland tropical forest in Panama.

The relatively large decline of TC, available P, and exchangeable K, Ca, Mg, and Na concentrations was mostly recorded in soil from the 16-year-old mining site in Lati than in soil from the 5-year-old mining site in Sambarata. This implied that mining, particularly long-term mining, severely lowered soil quality. Consistent with our results, Ghose (2002) reported reductions of CEC and C, N, P, K, Ca, Mg, and Na concentrations with increasing age of overburden from 1-10 years due to erosion in mined soil dump. On the other hand, Maharana and Patel (2013) reported increases in soil C, N, and P concentrations with increasing age of overburden from 0-10 years. The gradual establishment of vegetation cover on the overburden accounted for those results. Our results underscored the urgent need for rehabilitation to improve soil chemical properties at the post coal mining sites in East Kalimantan, Indonesia.

Impact of coal mining on plant growth, shoot nutrient concentration, and shoot nutrient content

Soil from post coal mining significantly showed the plant growth inhibition as expected. However, we found no symptoms linked to toxic metals, such as necrosis, bronze spotting, or chlorosis, in leaves. To clarify the reasons for plant growth inhibition, shoot nutrient concentrations were measured. The depletion of K and Mg concentrations in soil from the post coal mining sites could have resulted in the decreases in shoot K and Mg concentrations. By contrast, shoot P concentration was higher in *S. bicolor* grown in soil from the post coal mining site than in that grown in soil from the natural forest in both Lati and Sambarata. This could be explained by the dilution effects of high shoot dry weight production. As shown in Fig. 1, shoot dry weight was lower in *S. bicolor* grown in soil from the natural forest. Low shoot N, P, and K concentrations with high biomass were previously observed in lodgepole pine at an oil shale post mining site (Kuznetsova *et al.*, 2011). High shoot Mg, K, and Ca concentrations with low

biomass were also observed in *Pinus sylvestris* grown at a brown coal mining site (Baumann *et al.* 2006). Based on the classification of sorghum nutrient concentration (Reuter and Robinson, 1986), shoot P concentration in this study is categorized as deficient as this value is <20 mgP.g⁻¹ in *S. bicolor* grown in soil from both natural forest and post coal mining sites in Lathi and Sambarata. In contrast, shoot K, Ca, Mg, and Fe concentrations are categorized as adequate. Our results indicate that P is the major plant growth limiting factor. Low shoot N, P, K, Ca, and Mg concentrations were also observed in grass grown in coal spoils in the Lusatian region, Germany (Nada *et al.*, 2011).

Shoot dry weight had significant positive correlations with soil chemical properties in both Lati and Sambarata (Table 4). The very low N, P, K, Ca, Mg, and TC concentrations in soil from the post coal mining site decreased plant growth. Positive correlations between shoot dry weight and all the soil chemical properties were observed in plants grown in soil from Lati post coal mining site. The lower soil chemical properties in post coal mining site resulted in lower shoot dry weight of S. bicolor in comparison to this plant grown in soil from natural forest. It indicates that all the soil chemical properties are the limiting factors of plant growth at Lati post coal mining site. In contrast, only TN, TC, Mg, and Ca concentrations were positively correlated with shoot dry weight in plants grown in soil from Sambarata post coal mining site. The fact that the mining site in Sambarata area is only five years old may explain the low impact of mining, as shown by the lack of significant correlation of some soil chemical properties with shoot dry weight. However, as available P and K showed positive correlations with shoot dry weight, they are the potential limiting factors of plant growth in long-term mining. The results imply that low soil fertility, which is reflected by the low nutrient availability in soil from the post coal mining site, is influenced by mining period. The low soil nutrient availability suppresses nutrient uptake, thereby limiting plant growth. Consistent with our results, positive correlations between soil N and P and total biomass were observed in Alnus sinuata and Anaphalis margaritacea grown in copper mine tailings (Kramer et al., 2000). It is reported that the application of P improved growth of Oryza sativa, Triticum aestivum, and Brassica chinensis in a reclaimed coal mine in China (Chen et al., 1998), and the application of N, P, and K improved growth of 13 tropical tree species in coal mine spoils in India (Jha, 1992). The reduction of soil Ca, Mg, and K concentrations was also reported limited plant growth in lignite mine soil in Spain (Monterroso *et al.*, 1998).

To summarize, it was clarified that the inhibition of *S. bicolor* growth in soil from post coal mining sites is caused by multiple nutrient deficiencies. Generally, the low availability of soil nutrients inhibited nutrient uptake by and growth of plants grown in soil from the post coal mining sites in both Lati and Sambarata. This plant growth inhibition was more pronounced in soil from the 16-year-old mining site in Lati than in soil from the 5-year-old mining site in Sambarata (Tables 4). This result indicates potential reforestation problems at post coal mining sites due to the reduction of soil fertility.

In conclusion, our results show that coal mining changes the soil chemical properties into under limit and exhibit the plant growth inhibition. This study also shows that the duration of ecosystem disturbance by coal mining is related to soil fertility. This is the main reason for the inhibition of plant growth in soil from the post coal mining sites in humid tropical Indonesia. Nevertheless, the present results do not represent the diverse site conditions in Indonesia, but illustrate the relationship between plant growth and soil under mining impact. To overcome this soil degradation problem at post coal mining sites, rehabilitation is urgently needed. Application of soil organic amendment, K, Ca, Mg, P, and plant growth-promoting microorganisms may improve nutrient supply and soil fertility for reforestation.

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Area			Ex	changeable cation			
	Total Nitrogen (%)	Avail. P ⁺ (mg P ₂ O ₅ kg ⁻¹)	Total Carbon (%)	K	Ca (cmol _c kg ⁻¹)	Mg	CEC (cmol _c kg ⁻¹)
Lati: Shoot dry weight (mg pot ⁻¹) Sambarata:	+0.821**	+0.841***	+0.799**	+0.727**	+0.738**	+0.954***	+0.823***
Shoot dry weight (mg pot-1)	$+0.644^{*}$	+0.521ns	+0.679*	+0.536ns	+0.655*	+0.622*	+0.479ns
[†] Avail. P, available phosphate; []] cance; *** 0.1% level of significa	K, potassium; Ca, nce; ns: not signifi	calcium; Mg, magne cant.	sium; CEC, catio	n exchange capaci	ty. * 5% level of sig	gnificance; ** 1% l	evel of signifi-

Table 4. Correlations between shoot dry weight of *Sorghum bicolor* and chemical properties of soil from natural forest and post coal mining site in Lati and

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References

- Abreu, M.M., Santos, E.S., Ferreira, M. and Magalhaes, M.C.F. 2012. *Cistus salviifolius*, a promising species for mine wastes remediation. J. Geochem. Explor. 113: 86-93.
- Alvarez-Clare, S. and Mack, M.C. 2011. Influence of precipitation on soil and foliar nutrients across nine Costa Rican forests. *Biotropica*. 4 : 433-441.
- Baumann, K., Rumpelt, A., Schneider, B.U., Marschner, P. and Huttl, R.F. 2006: Seedling biomass and element content of *Pinus sylvestris* and *Pinus nigra* grown in sandy substrate with lignite. *Geoderma*. 136: 573-378.
- Bergmann, C., Stuhrmann, M. and Zech, W. 1994. Site factors, foliar nutrient levels, and growth of *Cordia alliodora* plantations in the humid lowlands of Northern Costa Rica. *Plant Soil*. 166 : 193-202.
- Bian, Z., Inyang, H., Daniels, J.L., Otto, F. and Struthers, S. 2010. Environmental issue from coal mining and their solutions. *Min. Sci. Technol.* 20 : 0215-0223.
- Braghina, C., Peptenatu, D., Constantinescu, S., Pintilii, R.D. and Draghici, C. 2010. The pressure exerted on the natural environment in the open pit exploitation areas in Oltenia. *Carpath. J. Earth Env.* 5 : 33-40.
- Burger, J.A. and Zipper, C.E. 2002. How to restore forest on surface-mined land. *Virginia Cooperative Extension*. Publication 460-123. Virginia.
- Carroll, C. and Tucker, A. 2000. Effects of pasture cover on soil erosion and water quality on central Queensland coal mine rehabilitation. *Trop. Grassland*. 34: 254-262.
- Chen, H., Zheng, C. and Zhu, Y. 1998. Phosphorus: a limiting factor for restoration of soil fertility in a newly reclaimed coal mined site in Xuzhou, China. *Land Degrad Dev.* 9 : 115-121.
- Directorate General of Minerals and Coal, 2012: Mineral of Energy and Mineral Resources, Indonesia. http:// www.djmbp.esdm.go.id/index_dbm.php (May, 2013).
- Dowarah, J., Boruah, H.P.D., Gogoi, J., Pathak, N., Saikia, N. and Handique, A.K. 2009. Eco-restoration of a high-sulphur coal mine overburden dumping site in northeast India: A study case. *J. Earth Syst. Sci.* 118: 597-608.
- Flores, C.I., Clark, R.B. and Gourley, L.M. 1988. Growth and yield traits of sorghum grown on acid soil at varied aluminum saturation. *Plant Soil.* 106: 49-57.
- Flores, C.I., Gourley, L.M., Pedersen, J.F. and Clark, R.B. 1991. Inheritance of acid soil tolerance in sorghum (*Sorghum bicolor*) grown on an Ultisol. In: RJ Wright et al. (Ed.) *Plant-Soil Interactions at Low pH*, pp. 1081-1093. Kluwer Academic Publishers, The Netherlands.

- Food and Agriculture Organization 2009: Indonesia outlook study. http://www.fao.org/docrep/014/ am608e/am608e00.pdf (February, 2012).
- Ghose, M.L. 2002. Effects of erosion on some properties of soil within the top 0.2 m of storage dumps. *Land Contam. Reclam.* 10 : 107-114.
- Ghose, M.K. 2004. Effect of open cast mining on soil fertility. J. Sci. Ind. Res. India. 63 : 1006-1009.
- Ghose, M.K. and Majee, S.R. 1998. Assessment of dust generation due to opencast coal mining-an Indian case study. *Environ. Monit. Assess.* 61 : 255-263.
- Hazelton, P. and Murphy, B. 2007. *Interpreting Soil Test Results: What Do All the Numbers Mean?* (2nd Ed.). CSIRO publishing, Australia.
- Jha, A.K. 1992. Evaluation of coal mine spoil as medium for plant growth in a dry tropical environment, India. *Indian For.* 118 : 909-916.
- Jha, A.K. and Singh, J.S. 1991. Spoil characteristics and vegetation development of an age series of mine spoils in a dry tropical environment. *Vegetation*. 97: 63-76.
- Juwarkar, A.A. and Jambhulkar, H.P. 2008. Phytoremediation of coal mine spoil dump through integrated biotechnological approach. *Bioresource Technol.* 99(11) : 4732-4741.
- Kien, C.N., Noi, N.V., Son, L.T., Ngoc, H.M., Tanaka, S., Nishina, T. and Iwasaki, K. 2010. Heavy metal contamination of agricultural soils around a chromite mine in Vietnam. *Soil Sci. Plant Nutr.* 56 : 344-356.
- Kramer, P.A., Zabowski, D., Scherer, G. and Everett, R.L. 2000. Native plant restoration of copper mine tailings: I. Substrate effect on growth and nutritional status in a greenhouse study. *J. Environ. Qual.* 29: 1762-1769.
- Kuznetsova, T., Tilk, M., Parn, H., Lukjanova, A. and Mandre, M. 2011. Growth, aboveground biomass, and nutrient concentration of young Scot pine and lodgepole pine in oil shale post-mining landscape in Estonia. *Environ. Monit. Assess.* 183 : 341-350.
- Maharana, J.K. and Patel, A.K. 2013. Physico-chemical characterization and mine soil genesis in age series coal mine overburden spoil in chronosequence in a dry tropical environment. *J. Phylogenet. Evol. Biol.* 1: 101. doi:10.4172/jpgeb.1000101
- Makineci, E., Gungor, B.S. and Kumbasli, M. 2011. Natural plant revegetation on reclaimed coal mine landscape in Agacli-Istanbul. *Afr. J. Biotechnol.* 10 (16): 3248-3259.
- Monjezi, M., Shahriar, K. and Dehghani, H. 2009. Environmental impact assessment of open pit mining in Iran. *Environ. Geol.* 58 : 205-216.
- Monterroso, C., Macias, F., Gil Bueno, A. and Val Caballero, C. 1998. Evaluation of the land reclamation project at the AS Pontes mine (NW Spain) in relation to the suitability of the soil for plant growth. *Land Degrad. Dev.* 9 : 441-451.
- Nada, W.M., Leon van Rensburg, Claassens, S. and

Blumenstein, O. 2011. Effect of vermicompost on soil and plant properties of coal spoil in the Lusatian Region (Eastern Germany). *Communications in Soil Science and Plant Analysis*. 42(16) : 1945-1957. www.tandfonline.com/toc/lcss20/current (June, 2013).

- Ohta, S. and Effendi, S. 1992. Ultisols of "Lowland *Dipterocarp* Forest" in East Kalimantan, Indonesia. II. Status of Carbon, Nitrogen, and Phosphorus. *Soil Sci. Plant Nutr.* 38(2) : 207-216.
- Olsen, S.R. and Sommers, L.E. 1982. Phosphorus. In: Page AL (Ed). *Methods of Soil Analysis. Part 2. Chemical and Microbiological Properties*, pp. 403-430. Soil Science Society of America, Madison.
- Peplow, D. and Edmonds, R. 2005. The effects of mine waste contamination at multiple levels of biological organization. *Ecological Engineering*. 24 : 101-119.
- Resosudarmo, B.P., Resosudarmo, I.A.P., Sarosa, W. and Subiman, N.L. 2009: Socioeconomic conflicts in Indonesia's mining industry. *In* Cronin, R (Ed.), *Exploiting Natural Resources: Growth, Instability, and Conflict in the Middle East and Asia,* pp. 33-46. The Henry L. Stimson Center, Washington DC.
- Reuter, D.J. and Robinson, J.B. 1986. *Plant analysis: An interpretation manual*. Inkata Press, Sydney, Australia.
- Rivera-Becerril, F., Lucia, V. and Juarez-Vazquez, 2013. Impact of manganese mining activity on the environment: interaction among soil, plants, and arbuscular mycorrhiza. *Archives of Environmental Contamination and Toxicology*. 64 : 219-227.
- Sadhu, K., Adhikari, K. and Gangopadhyay, A. 2012. Effect of mine spoil on native soil of lower Gondwana coal fields: Raniganj coal mines areas, India. *Int. J. Environ. Sci.* 2(3) : 1645-1687.
- Santiago, L.S., Schuur, E.A.G. and Silvera, K. 2005. Nutrient cycling and plant-soil feedbacks along a precipitation gradient in lowland Panama. *J. Trop. Ecol.* 21: 461-470.
- Sheoran, V., Sheoran, A.S. and Poonia, P. 2010. Soil reclamation of abandoned mine land by revegetation: a review. Int. J. Soil, Sediment Water. 3, Iss. 2, Article 13.

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- Shi, X., Zhang, X., Chen, G., Chen, Y., Wang, L. and Shan, X. 2011. Seedling growth and metal accumulation of selected woody species in copper and lead/zinc mine tailings. J. Env. Sci. 23 : 266-274.
- Shrestha, R.K. and Lal, R. 2011. Changes in physical and chemical properties of soil after surface mining and reclamation. *Geoderma*. 161 : 168-176.
- Singh, R.S., Tripathi, N. and Chaulya, S.K. 2012. Ecology study of revegetated coal mine spoil of an Indian dry tropical ecosystem along an age gradient. *Biodegradation*. 23 : 837-849.
- Srivastava, S.C., Jha, A.K. and Singh, J.S. 1989. Changes with time in soil biomass C, N, and P of mine spoils in a dry tropical environment. *Can. J. Soil Sci.* 69: 849-855.
- Station of Meteorology of Tanjung Redeb, Berau Regency 2008: Climate. http://<u>www.beraukab.go.id</u> (March, 2013).
- Suryantoro, S. and Manaf, M.H. 2002. The Indonesian energy and mineral resources development and its environmental management to support sustainable national economic development. In The OECD Conference in Foreign Direct Investment and Environment in Mining Sector. Paris, France.
- Syahrial, E. Adam, R. Suharyanti. Ajiwihanto, N. Indarwati, R. R. F. Kurniawan, F. Kurniawan, A. and Suzanti, V. M. 2012. Handbook of energy & economic statistics of Indonesia. Center for Data and Information on Energy and Mineral Resources. Indonesia Ministry of Energy and Mineral Resources.
- Truog, E. 1930. The determination of the readily available phosphorus of soils. *J. Am. Soc. Agron.* 2 : 874-882.
- Vernimmen, R.R.E., Verhoef, H.A., Verstraten, J.M., Bruijnzeel, L.A., Klomp, N.S., Zoomer, H.R. and Wartenbergh, P.E. 2007. Nitrogen mineralization, nitrification and denitrification in Central Kalimantan, Indonesia. *Soil Biol. Biochem.* 39 : 2992-3003.
- Wong, M.H. 2003. Ecological restoration of mine degraded soils, with emphasis on metal contaminated soils. *Chemosphere*. 50 : 775-780.