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# Voltametric trace determination of Vanadium ions and their removal from industrial effluents by using biosorbent

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## ABSTRACT

Vanadium is extensively utilized in a variety of industries, including textiles, chemical manufacturing, metal plating, and many more. Pentavalent Vanadium is a highly toxic and carcinogenic substance when present in high concentrations. Therefore, eliminating it from industrial effluents before it can reach any water supply or land is critical. In the present research, the determination of trace V(V) by Differential pulse polarography (DPP) and its removal from industrial effluents by using Annona squamosa leaves as biosorbent has been analyzed. The mercury electrode is used as working electrode for DPP analysis. The reference and auxiliary electrodes were Ag/AgCl and platinum wire respectively. The biosorbent capacity of Annona squamosa leaves were analyzed by using different parameters like pH, concentration, contact time, biosorbent dosage, and temperature. The V(V) showed a sharp DP peak at -1.19V in presence of Ammonia-Ammonium phosphate buffer. The maximum V(V) adsorption was 8.1mg/g and 85% with respect to contact time and concerning temperature respectively. The present study concluded that the determination of V(V) by DPP are in good agreement with that of results obtained from UV-VIS technique. Further the biosorbent so prepared from Annona squamosa leaves is likely to be a cost-effective, economically feasible, and easily available material for removing V(V)from industrial effluents.

Key words : Vanadium, Differential pulse polarography, Annona squamosa, and Biosorbent.

## Introduction

Industrialization has resulted in a huge increase in the emission of various pollutants into the environment. Heavy metal contamination in wastewater from industrial processes such as mining, smelting, and battery manufacture is one of the world's most critical environmental issues (Gosar, 2004). Heavy metals are one of the most dangerous sorts of pollution in water. Cu, Co, Se, V, Cr, Fe, Ni, Cd, Pb, As, Hg, Zn, and other poisonous metal ions are the most dangerous to humans and other forms of life. Because of their toxicity, bioaccumulation potential, and natural persistence, these heavy metals are of particular concern (Wang, 2011; Jai *et al.*, 2007; Bailey *et al.*, 1999). Vanadium is abundant in the environment and plays a biological role in most living things. But the existence of Vanadium in the environment above a certain limit is toxic to human health. Vanadium is a persistent environmental pollutant that is released into the environment by numerous industries that work with alloy steel (Stasko *et al.*, 2006; Tang *et al.*, 2008; Gerke *et al.*, 2010 and Liber *et al.*, 2011). They are a serious issue in terms of environmental and human health implications due to their toxicity and build-up in the food chain. The International Agency for Research on Cancer categorized vanadium pentoxide as a probable carcinogen. Vanadium is currently on the USEPA's (the United States Environmental Protection Agency) Drinking Water Contaminant Candidate List (CCL3) because of its carcinogenic potential (Naeem *et al.*, 2007 and Leiviska *et al.*, 2017). Vanadium concentrations in drinking water can range from 0.2 to 100  $\mu$ g/l, with normal levels between 1 and 6  $\mu$ g/l(Ghazvini *et al.*, 2009 and Huang *et al.*, 2015).

The detection of Vanadium ions contamination in wastewater is a critical step in protecting human and environmental health. The need for low-cost, effective methods to remove Vanadium metal ions from wastewater is critical.

Chemical precipitation, ion exchange, chemical oxidation, reduction, reverse osmosis, ultrafiltration, electrodialysis, and adsorption are some of the heavy metal removal processes used (Wang, 2011). Determination of Lanthanum in industrial samples was also achieved by DPP (Sharma *et al.*, 2015). Differential pulse polarography was employed to investigate lead and cadmium in synthetic and natural hair dyes (Majda *et al.*, 2021).

Natural biosorbents are made from living tissue and are quite inexpensive. Therefore, while choosing an adsorbent for wastewater remediation, budget analysis is a critical factor to consider. The price of the adsorption process is determined by the adsorbent's price. Commercial activated carbon, for example, costs Rs. 500/kg; however, bio sorbents cost between Rs. 4.4 to Rs. 36.89/kg, which is significantly less than commercialized adsorbents. (Gupta and Babu, 2008). For heavy metal removal, bio-adsorption techniques are widely utilized in wastewater treatment. Biosorbents have the advantages of being inexpensive, readily available, producing no sludge, being regenerated, and having technological feasibility, engineering utility, and affinity for the elimination of toxic metals.

In this study, the determination of Vanadium in the industrial wastewater was achieved by differential pulse polarography, and the results obtained were compared with UV-VIS spectroscopy. The absence of any chelating agent, simplicity, low cost, high sensitivity, sufficient selectivity, precision, and accuracy are the most significant advantages of the suggested method. Further in our investigation Annona Squamosa leaves powder has been studied to see if they might be used in useful ways, particularly in wastewater treatment. Their porosity nature makes them an appealing material for use as an adsorbent. Eco. Env. & Cons. 28 (October Suppl. Issue) : 2022

The major goal of this research was to develop a practical voltammetric technology with a straightforward approach for determining vanadium at micro levels and its removal with a cost-effective and pollution-free biosorbent.

## **Materials and Methods**

#### Instrumentation

Model 174-A polarographic analyser with X-Y recorder (Model RE 0074) and drop timer (Model 174/ 70). Polarographic recordings were made with (EG&G, USA). The following were the instrumental settings for DPP: The working electrode was a falling mercury electrode (DME). 50 mV pulse duration, 50 mV pulse amplitude Pulse time is 57 milliseconds, and the scan rate is 0.5 seconds. 5mVsec. The reference and auxiliary electrodes were Ag/ AgCl and platinum wire, respectively.

The findings of polarographic investigations with a wavelength range of 190-900 nm were compared using a UV-VIS spectrophotometer (Model, 108) from Systronics, India. The light source and detector were a tungsten-halogen deuterium lamp and a broad range photomultiplier, respectively. The resolution spectral bandwidth was 0.5 nm.

For adsorption studies weighing machine, pH meter, Jasco made FT-IR spectrophotometer were used.

#### **Sample Preparation**

Industrial effluents samples were gathered from metal and alloy industries in Jodhpur's industrial zones. To separate any suspended particle materials, these were filtered. The water samples were preconcentrated to 110 of their original volumes by boiling 100 ml of this aliquot to obtain detectable amounts of metals. The contents were heated until the solution fumed, removing biological and organic matrices. I ml of the oxidizing mixture (HNO<sub>3</sub>:  $H_2SO_4$ ; 5ml: 1ml) was added, and the contents were heated until the solution fumed, removing biological and organic and organic matrices.

For the preparation of the biosorbent Annona, Squamosa leaves were collected from the Udaipur district of Rajasthan. The leaves were washed thoroughly to remove dirt, dried in shade, and grounded with mortar. Approx. 50 g of grounded leaves were treated with 100 ml of HNO<sub>3</sub> and deionized water to prepare the biosorbent.

#### Chemicals

All the chemicalsused were of analytical grade purity. The stock solutions of vanadium wereprepared from Vanadium pentoxide ( $V_2O_5$ ), Batch No.– 10160, Riedel-dehaenag-Seelze-hanmover.

All the sample solutions were made in purified water obtained from Elix (Mili-CT Advantage A 10) of Millipore, SAS, France.

## **Biosorption studies**

The adsorption experiment was carried out at various pH levels, interaction times, initial V (V) concentrations, and biosorbent dosages. Before and after V (V) uptake, the biosorbent was characterized using FTIR. The percentage of V(V) ions removed from industrial effluents were calculated by the given equation: -

$$\begin{array}{c} \mbox{Percentage removal of V(V) ions = } & \hline C_i - C_f \\ \hline C_i \end{array} \\ \hline \end{array} \begin{array}{c} \mbox{100} \end{array}$$

 $C_i$ - initial concentration of biosorbent,  $C_f$ -Final concentration of the biosorbents.

The amount of V (V) ion absorbed by biosorbent  $(q_e)$  at equilibrium was determined using the given formula-

 $q_e = \frac{(C_i - C_f) X V}{1000 X w}$ 

where  $C_i$  is the initial metal ion concentration (mg/L),  $C_f$  is the final concentration at equilibrium (mg/L), and w is the mass of the biosorbent (g), and V is the volume of the solution (L).

# **Results and Discussion**

## Differential pulse polarographic studies

#### **Choice of Supporting Electrolyte**

Preliminary experiments on vanadium (V) electroreduction on DME suggested that an ammonia-ammonium phosphate buffer medium might be

Table 1. Polarographic characteristics of vanadium (V)



Fig. 1. DC Polarogram of vanadium (V)

- (a) Blank solution of ammonia-ammonium phosphate buffer (pH 3.2)
- (b) 1.0 x 10<sup>-5</sup>M V (V) in ammonia-ammonium phosphate buffer (pH 3.2)

suitable. With a half–wave potential of –1.19 V, a single DC polarographic wave was obtained (Fig. 1.1). Table 1 shows the polarographic features of vanadium (V).

Vanadium (V) showed a sharp DP peak at -1.19 V in the presence of an ammonia-ammonium phosphate buffer, according to DPP experiments. Figure 2 depicts the situation. Peak current versus vanadium concentration was found to be linear between 0.008 and 35 ppm.

Figure 3 shows DP polarograms recorded at various vanadium (V) concentrations. Figure 4 shows a calibration curve.

Calibration curve characteristics:

Slope	=	0.0624
Intercept	=	0.0837
Coefficient of correlation (r)	=	1.0775

## Limit of Quantification

Under these conditions, the minimum concentration of vanadium (V) that could be calculated was 0.008 g/ml.

Supporting electrolyte	-E <sub>1/2</sub> (V)	Ι <sub>d</sub> (μΑ)
1M hydrochloric acid	1.45	1.3
1M sulphuric acid	0.91	0.5
1M ammonia	1.35	2.4
Sodium citrate buffer ( $pH = 8.1$ )	1.37	2.9
Ammonium oxalate buffer ( $pH = 9.2$ )	1.79	2.6
Ammonia–ammonium chloride buffer ( $pH = 8.4$ )	1.28	1.4
0.5 M ammonia–ammonium acetate buffer	1.35	2.2
Ammonia-ammonium phosphate buffer (pH= 3.2)	1.19	3.1
	Supporting electrolyte 1M hydrochloric acid 1M sulphuric acid 1M ammonia Sodium citrate buffer (pH = 8.1) Ammonium oxalate buffer (pH = 9.2) Ammonia–ammonium chloride buffer (pH = 8.4) 0.5 M ammonia–ammonium acetate buffer Ammonia-ammonium phosphate buffer (pH= 3.2)	Supporting electrolyte $-E_{1/2}$ (V)1M hydrochloric acid1.451M sulphuric acid0.911M ammonia1.35Sodium citrate buffer (pH = 8.1)1.37Ammonium oxalate buffer (pH = 9.2)1.79Ammonia-ammonium chloride buffer (pH = 8.4)1.280.5 M ammonia-ammonium acetate buffer1.35Ammonia-ammonium phosphate buffer (pH = 3.2)1.19



Fig. 2. DP Polarogram V (V)

- (a) Blank solution of ammonia-ammonium phosphate buffer (pH 3.3)
- (b) 24.5 ppm vanadium (V) in ammonia-ammonium phosphate buffer (pH 3.3)



Fig. 3. DP Polarograms of vanadium (V) at different concentrations in ammonia-ammonium phosphate buffer (pH 3.3)



Fig. 4. Calibration curve of vanadium (V) concentration Vs. peak height

#### Interferences

During the DPP determination of vanadium, the interference of copper, lead, cadmium, and zinc was observed. Industrial wastewaters are more likely to contain these metal ions. In an ammonia-ammonium phosphate buffer, Cu (II), Pb (II), Cd (II), and Zn (II) produced distinct DP peaks. As a result, these metal ions had no effect. Figure 1.5 depicts the situation. Table 2 shows the peak potentials.

**Table 2.** Peak potential of metal ions in ammonia-ammo-nium phosphate buffer (pH 3.3)

S.No.	Interfering ion	–Ep (V) vs SCE
1.	Copper (II)	0.03
2.	Lead (II)	0.47
3.	Cadmium (II)	0.66
4.	Zinc (II)	1.14
5.	Vanadium (V)	1.25



Fig. 5. DP Polarogram of V (V) in presence of Cu (II), Pb (II), Cd (II), and Zn (II) in ammonia-ammonium phosphate buffer (pH 3.3) Cu (II)= 11 ppm; Pb (II)= 6 ppm; Cd (II)= 8 ppm; Zn (II)= 11 ppm and V (V)= 16 ppm.

#### **Precision and Accuracy**

By evaluating vanadium (V) in ammonia-ammonium phosphate buffer (pH 3.3), where its concentration was preserved at 0.5 ppm, DPP observations were examined for precision and accuracy in the form of standard variance. Table 3 summarizes the findings.

#### **Determination of Vanadium in Diverse Matrices**

The vanadium concentration in industrial waste streams and dietary supplements was determined by DP polarographic reduction of vanadium (V) in ammonia-ammonium phosphate buffer (pH 3.3).

 Table 3. Precision and accuracy of vanadium (V) determination by DPP

S. No.	V (V) concn. (ppm) Present Determined		S.D.(±) (%)	R.S.D. Error (%)	Percent- age
1. 2. 3. 4. 5.	0.5 0.5 0.5 0.5 0.5	0.481 0.489 0.493 0.497 0.479	0.0077	1.578	2.44

Sampling sites	Van	Vanadium content.(µg/ml)			C.V.(%)
	Min	Max	Ave.		
Textile industry 1	1.63	1.68	1.65	0.03	1.82
Textile industry 2	1.15	1.24	1.19	0.06	5.04
Alloy industry	1.95	2.02	1.98	0.05	2.52

Table 4. Concentration of vanadium in industrial wastewater

N = 5, Number of samples

= Jodhpur industrial area

The corresponding data were shown in Table 4 and 5.

 Table 5. Concentration of vanadium in the mineral supplement

S. No.	Vanadium concentration (µg/ml)*	Ave.	S.D.(±)	R.S.D. (%)	Error (%)
1. 2. 3. 4. 5.	3.64 3.71 3.69 3.73 3.66	3.69	0.04	1.08	± 2.5

\*cited value 3.6 µg/ml vanadium

#### Comparison

To demonstrate the validity of DPP measurements, the findings of vanadium (V) determination were compared using UV–VIS spectrophotometry. Table 6 provides comparative data.

#### **Biosorption studies of V(V) ions**

## Effect of pH on removal of V(V) ions

The pH value is a key component in influencing the adsorption process, as the surface charge of the adsorbents, the degree of ionization, and the species of the adsorbate are all affected by pH. The electrostatic interaction, or attraction between adsorbent and adsorbate, is the primary driving force behind

 Table 6. Determination of vanadium by DPP and UV–

 VIS spectrophotometric method

S.	Sample Vanadium conc		
No.		DPP	UV–VIS
1.	Wastewater		
	Textile industry 1	1.65	1.54
	Textile industry 2	1.19	1.13
	Alloy industry	1.98	1.86
2.	Mineral Supplement	3.69	3.59

metal ion adsorption. Heavy metal adsorption was shown to be higher when the interaction was high (Krishnaand Sreeb, 2013).

In a pH range of 1-8, the influence of pH solution was investigated using 5 ml V (V) 15 mg/l, 0.25 g biosorbent, and an interaction period of 10 minutes with a stirrer speed of 150 rpm. Figure 6 shows the outcome. With increasing pH and pressure, adsorption capacity increased from 1.2 mg/g to 1.9 mg/g. Then it reduced. Since pH affects metal ion adsorption, it's one of the most important factors to consider that directly influences metal solubility or the degree of dissociation of organic functional groups on the sorbent's membrane. Normally alkaline circumstances will reduce the rate of adsorption of metal because a higher pH produces fewer protons, which reduces the competition between proton and heavy metal ions (Homaidan et al., 2014). At low pH, there is a partial increase in the metal adsorption rate due to competition between H<sup>+</sup> ions with metal ions for exchange sites available on the surface of the biosorbent (Khan et al., 2013).

#### Effect of initial concentration of V (V) ions

At pH 7, the effect of V (V) concentration was tested



**Fig. 6.** Effect of pH on biosorption of V(V) by Annona squamosa biosorbent.

at various doses ranging from 100 to 500 mg/l and keeping the biosorbent concentration fixed. The optimum adsorption capacity of the Annona Squamosa biosorbent was attained at V(V) 400 mg/l with a value of 6.9 mg/g, as shown in Fig. 7. With the increase in the concentration of V(V) ions the Biosorption capacity has increased and then reduced after attaining the equilibrium. It was found that with the increase in V(V) ion concentration, V(V) adsorption by the biosorbent increases. All V(V) ions interact with the binding sites at lower concentrations, resulting in maximal adsorption at low concentrations. However, at higher concentrations, V(V) ions will seek out free binding sites available on the biosorbent, but adsorption will decrease due to a lack of binding sites for complexation. (Jevaseelaand Gupta, 2016). The percentage biosorption of V(V) ions decreases with increasing concentration after attaining equilibrium is shown in Fig. 7.



**Fig. 7.** Effect of initial concentration on biosorption of V(V) by Annona squamosa biosorbent.

#### Effect of biosorbent dosage on removal of V(V)ions

The amount of *Annona squamosa* biosorbent utilized in this investigation ranged between 0.2 and 1 g. Figure 8 depicts the outcome. The biosorption capacity was improved at V(V) low doses of biosorbent mass. With a biosorbent dose of 0.4 g, the optimum sorption capacity for Annona Squamosa L leaves powder was 6.5 mg/g. As the biosorbent dosage was increased, the sorption capacity of the biosorbent dropped. This could be explained by the fact that the lower the biosorbent ratio and the metal retained by a sorbent at a fixed concentration in suspension (Homaidan *et al.*, 2014). The absorption of metal ions increases as the concentration of biosorbent de-



**Fig. 8.** Effect of biosorbent (*Annona squamosa*) dosage on biosorption of V(V).

creases (Raoand Prabhakar, 2011).

Effect of contact time on the removal of V(V) ion-The effect of contact time on V(V) biosorption by Annona squamosa biosorbent was investigated with contact times ranging from 30 to 150 minutes. The maximum biosorption capacity of Annona Squamosa biosorbent was observed at contact time 90 min with a value of 8.1 mg/g, as shown in Fig. 9. Biosorption capacity increased for the first 90 minutes of contact time and thereafter declined. The contact time is a critical parameter for the efficient elimination of metal ions in an aqueous environment. The decrease in the coefficient of mass transfer of the diffusion-controlled reaction between the sorbent and the metal ions can explain the increased sorption of metal ions with increasing contact time. (Sayed et al., 2010).

There are two steps in the adsorption process. The initial stage of adsorption was fast, but the sec-



Fig. 9. Effect of contact time on the removal of V(V) by Annona squamosa biosorbent.

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ond stage was slower. This can be explained as follows: in the beginning, a significant number of vacant sites may be present, resulting in rapid absorption; however, as the vacant sites are depleted, adsorption declines. As a result of repulsion between the solute molecules and the bulk phase (Qaiser *et al.*, 2009).

## Effect of Temperature on the removal of V(V) ions

At different temperatures  $(25^{\circ}C, 30^{\circ}C, 35^{\circ}C, 40^{\circ}C, 45^{\circ}C \text{ and } 50^{\circ}C)$ , the adsorption of V(V) ions was investigated. The percentage of removal of V(V) ions increases with increasing temperature, reaching a high of 85% at 50°C given in Fig 10. The readings showed that the adsorption phenomenon is a chemisorption mechanism.



**Fig. 10.** Effect of temperature on the removal of V(V) by Annona squamosa biosorbent.

**FT-IR Analysis-**The FT-IR spectra of V(V) ions before and after adsorption are shown in the Fig 11. The peak shown between 3300 cm<sup>-1</sup> and 3450 cm<sup>-1</sup> is

of N-H stretching from -NH-OH group. The small sharp peak at 3250cm<sup>-1</sup> is due to O-H stretching also from -NH-OH group. The peak shown at 2821cm<sup>-1</sup> is due to N-H stretching in Primary amine. The shifting in the peak position in graph(a) and graph (b) clearly indicates that the adsorption takes place between V(V) ions and biosorbent so prepared from Annona squamosa leaves.

## Conclusion

An improved analytical technique for determining vanadium at micro levels (sensitivity, 0.008 g/ml) has been created as a result of comprehensive DPP investigations. The DPP results were found to be pretty close to those obtained using the UV-VIS technique. Hopefully, the approach devised will be useful in research labs, metal companies, and pharmaceuticals for vanadium analysis. Further, the leaves of Annona squamosa due to their easy availability, low cost, and good adsorption capacity proved as promising biosorbents for the removal of V(V) ions from industrial wastewater. There is a need to be carried out further research to make the approach commercially viable to remove V(V) ions from industrial effluents.

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Fig. 11. FTIR spectrum of Annona squamosa biosorbent (a) before adsorption of V(V) ions (b) after adsorption of V(V) ions.

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## References

- Bailey, S.E., Olin, T.J., Bricka, R.M. and Adrian, D.D. 1999. A review of potentially low-cost sorbents for heavy metals. *Water Research*. 23(11) : 2469-2479.
- Gerke, T.L., Scheckel, K.G. and Maynard, J.B. 2010. Speciation and distribution of vanadium in drinking water iron pipe corrosion by-products. *The Science of the total Environment*. 408 : 5845-5853.
- Ghazvini, M.P.T. and Ghorbanzadeh, S.G. 2009. Bioresource technology effect of salinity on vanadate biosorption by *Halomonas* sp. GT-83: Preliminary investigation on biosorption by micro-PIXE technique. *Bioresour. Technol.* 100 : 2361-2368.
- Gosar, M. 2004. Environmental impacts of metal mining. Mater. Geoenviron. 51 : 2097-2107.
- Gupta, S. and Babu, B.V. 2008. Economic feasibility analysis of low cost adsorbents for the removal of Cr (VI) from waste water. In: *Proceedings of International Convention on Water Resources Development and Management (ICWRDM), BITS Pilani*.23-26.
- Homaidan, AA.AL., Al-Houri, H.J., Al-Hazzani, A.A., Elgaaly, G. and Moubayed, N.S. 2014. Biosorption of copper ions from aqueous solutions by *Spirulina platensis* biomass. *Arabian Journal of Chemistry*. 7:57-62.
- Huang, J., Huang, F., Evans, L. and Glasauer, S. 2015. Vanadium: Global (bio) geochemistry. *Chem. Geol.* 417: 68–89.
- Jai, P.H., Wook, J.S., Kyu, Y.J., Gil, K.B.L. and Mok, L.S. 2007. Removal of heavy metals using waste eggshell. *Journal of Environmental Sciences*. 19 : 1436-1441.
- Jeyaseelan, C. and Gupta, A. 2016. Green Tea Leaves as a Natural Adsorbent for the Removal of Cr(VI) From Aqueous Solutions. *Air, Soil and Water Research.* 9 : 13-19.
- Khan, S., Farooq, A., Danish, M.I. and Zeb, A. 2013. Biosorption of Copper(II) From Aqueous Solution Using Citrus sinensis Peel and Wood Sawdust: Utilization In Purification of Drinking And Water. *IJRRAS.* 16 (2) : 297-306.
- Krishnaa, D. and Sreeb, R.P. 2013. Removal of Chromium from Aqueous Solution by Custard Apple (*Annona Squamosa*) Peel Powder as Adsorbent. *Int. J. Appl. Sci.*

Eng. 11 : 171-194.

- Leiviskaj, T., Khalid, M.K., Sarpola, A. and Tanskanen, J. 2017. Removal of vanadium from industrial wastewater using iron sorbents in batch and continuous flow pilot systems. *J. Environ. Manag.* 190 : 231–242.
- Liber, K., Doig, L.E. and White-Sobey, S.L. 2011. Toxicity of uranium, molybdenum, nickel, and arsenic to *Hyalellaazteca* and *Chironomus dilutus* in water-only and spiked-sediment toxicity tests. *Ecotoxicology and Environmental Safety*. 74 : 1171–1179.
- Majda, B., Chentoufi, M.A., Mojemmi, B., Benzeid, H., Toure, H.A., Cheikh, A., Rahali, Y., Idrissi, M.O.B., Draoui, M. and Bouatia, M. 2021. Determination of Lead and Cadmium in Synthetic and Natural Hair Dyes in Morocco Using Differential Pulse Polarography. *Portugaliae Electrochimica Acta*. 39(1): 37-44.
- Naeem, A., Westerhoff, P. and Mustafa, S. 2007. Vanadium removal by metal (hydr)oxide adsorbents. *Water Res.* 41:1596–1602.
- Qaiser, S., Saleemi, A.R. and Umar, M. 2009. Biosorption of lead from aqueous solution by *Ficus religiosa* leaves. Batch and column study. *J Hazard Mater*. 166: 998-1005.
- Rao, L.N. and Prabhakar, G. 2011. Removal of heavy metals by biosorption- an overall review. *Journal of En*gineering Research and Studies. 2 (4): 17-22.
- Sayed, GO.El., Dessouki, H.A. and Ibrahim, S.S. 2010. Biosorption of Ni (II) and Cd (II) ions from aqueous solutions onto rice straw. *Chem. Sci. J.* 9 : 1-11.
- Sharma, P., Bais, R. and Tiwari, M. 2015. Voltammetric trace determination of Lanthanum. *Inter. J. Adv. Res. in Science and Engg.* 4(7) : 311-316.
- Stasko, R., Adrian, H. and Adrian, A. 2006. Effect of nitrogen and vanadium on austenite grain growth kinetics of a low alloy steel. *Materials Characterization*. 56(4-5) : 340-347.
- Tang, Z. and Stumpf, W. 2008. The role of molybdenum additions and prior deformation on acicular ferrite formation in microalloyed Nb–Ti low-carbon linepipe steels. 59 : 717-728.
- Wang, F. and Fu, Q. 2011. Removal of heavy metal ions from wastewaters: A review. *Journal of Environmental Management.* 92 : 407-418.