

Evaluation of ecological status of the Shkumbini river based on Bio-chemical indicators

Lirika Kupe^{1*}, Aida Bani², Oljan Dervishi², Ilir Kristo² and Edmira Shahu³

¹*Department of Agronomy Science, Agricultural University of Tirana, Tirana, Albania,*

²*Department of Agro-Environment and Ecology, Agricultural University of Tirana, Tirana, Albania*

³*Department of Economy and Agrarian Policy, Agricultural University of Tirana, Tirana, Albania*

(Received 30 May, 2021; accepted 27 June, 2021)

ABSTRACT

Aims of this paper is the water quality assessment using bio-chemicals monitoring techniques, during July and September 2015, for the Shkumbini River, located in central Albania. Purpose is to provide more information about the ecological status and protection of these ecosystems. In this study from algae community, we used diatoms for Biomonitoring because they have a siliceous frustule. Three different indices (SI, TI_{DIA} and H') were used for biological evaluation. Based on diatom community structure, Shkumbini water quality was classified from mesotrophic to polytrophic, which shows a high status of eutrophication. Concentration of total phosphorus and nitrogen were used for the evaluation of trophic status for all sampling stations. This situation is due to the presence of inorganic and organic matter, originated from agricultural activity, restaurants, urban waste discharge and erosion by inhabitant centres.

Key words : Biomonitoring, Diatom, Nutrients, Freshwater,

Introduction

In the last decade high variation in the structure and function of river environment have been observed due to human activity and have also profound impacts upon the fluxes of nitrogen (N) and phosphorus (P) (Weigelhofer *et al.*, 2018). Variation of N and P caused by anthropogenic changes have strongly impacted the aquatic ecosystem too. Phosphorus and nitrogen are the primary drivers in eutrophication of aquatic ecosystems, where increased nutrient concentrations lead to increased primary productivity (Guignard *et al.*, 2017). N and P exported from different sources can have severe effects into the quality of fresh waters (Carpenter *et al.*, 1998; Correll, 1998). Water resources degradation due to the eutrophication can lead to depletion of their species component, as well as reduction of the facilities

or services provided by this watershed (Correll, 1998). Disposal of solid and domestic waste along the Shkumbini valley, especially in nearby urban areas, has an effect not only on the quality reduction but also to the aesthetic appearance (Paparisto *et al.*, 2010). Furthermore, the uncontrolled urban development and the lack of regulations on housing construction has led to improper disposal of construction waste, which is also a source of contamination for the rivers. The eutrophication phenomena along Shkumbini River are due to the high concentrations of nutrients (especially nitrogen and phosphorus), produced by human activities like, urban wastewater, industries, mining activities, gravel extraction, agriculture, and animal farming.

Water quality monitoring based on physico-chemical data usually do not offer a time integrated representation of the ecosystem's response to external

stresses, neither do they furnish valuable information about the impact of these variation on biological communities (Arle *et al.*, 2016). On the other hand, use of bio indicators can ensure an integrated data of measurement for the water quality as experienced by the aquatic biota and contribute with a helpful addition to chemical assessment based of water quality (Lavoie *et al.*, 2011; Jakovljević *et al.* 2016). The diatoms community is widely considered the first biological community to respond to eutrophication pressures and is the most direct of all the Biological Quality Elements of nutrient concentration in the water (John, 2002). Their use for biomonitoring has been developed to help determining quality of the aquatic environment trends and how that quality is influenced by waste treatment plant, or other anthropogenic activities (Chapman, 1996).

Diatoms respond to both water column nutrient concentration and habitat quality, and this has led to increasing interest in its use as a monitoring tool for running water like rivers. According to EN 13946 (2003) standard, samples of diatoms can also be collected from submerged hard surfaces, or submerged macrophytes. This Standard establishes also a method for the sampling and laboratory preparation of benthic diatoms for water quality assessments (STANDARD EN 14407 2004). Data produced by this method are suitable for production of water quality indices based on the relative dominance of taxa. With appropriate modifications, this method can also be utilized for the study of benthic diatoms in lakes. The Standard EN 14407 (2004) is applied for the identification and enumeration of the relative abundance of diatom taxa on preparation of permanent slides and interpretation of relevant data for the water quality assessment. Following European sectorial directives in the aquatic environmental field (Water Framework Directive and Environmental Quality Standards Directive) over recent decades, it pursues an integral approach for a uniform European water policy for the first time. Bio-monitoring methods are getting more and more intention in the last 100 years and are used in fresh waters to get information on the biota due to the impact from waste waters or fertilized agricultural lands (Whitton, 1995; Kupe *et al.*, 2006). Phytobenthic (diatoms) needs to be assessed in freshwater, with the assumption that these are particularly sensitive to nutrient and acidification also diatom community's changes are correlated with concentrations of both organic and inorganic materials (Sládeček, 1986; Kelly and

Whitton, 1995; Whitton, 1995; John, 2002; Prygiel *et al.*, 2002). Diatoms follow quickly environmental changes because they have a short lifecycle. They are easily sampled and preserved due to their siliceous cell wall, also allows water quality assessment on short or long-term changes. This has made them one of the preferred communities for biomonitoring all around the world (Cox, 1991; Kovacs *et al.*, 2006).

Based in the Water Framework Directive our aim was to evaluate diversity of phytobenthic, saprobe and the trophy level in Shkumbini River. The trophy level is based in three indices: Saprobic Index – SI (Rott *et al.*, 1999) which is correlated with organic matter; Trophic Diatom Index - TI_{DIA} (Rott *et al.*, 1997) is correlated with inorganic matter and Shannon Index (H') which show the diversity of species (Zelinka and Marvan, 1991). In the last decades, quality of waters has got larger attention in Albania, as well as their suitable management (Miho *et al.*, 2005; Kupe, 2006; Kupe *et al.*, 2008). At once is increasing the role of environmental monitoring, as a source of information for addressing environmental issues. The aim of this study was (i) to verify the degrees of organic pollution, (ii) usefulness of diatom indices in water quality assessments for Shkumbini River and (iii) to determine the dominant diatom taxa characteristic of the Shkumbini River.

Materials and Methods

Sampling sites

Shkumbini River is one of the most important rivers of Albania. It flows for 181 km in the center of Albania, with a drainage basin of 2444 km². It flows through the districts of Librazhd, Elbasan, Cerrik, Peqin, Kavaja and Lushnja, which are regions with high population density (Kabo, 1990-1991). Evaluation of water quality for the Shkumbini River, based on bio-chemical indicators was carried out in July and in September 2015. In this region, many effluents from local agriculture land, restaurant or other services are discharged directly to the river and furthermore untreated sewage from villages of Librazhd, Prrrenjas and Elbasan increase level of pollution. We have collected 10 samples in five stations (Karkavec - Prrrenjas, Dardhe - Librazhd, Labinot-Fushe, Paper and Peqin), (Fig. 1, Tab. 1), collected during July and September. The site stations were situated at different locations from human impact; the numbering is downstream the river, from the



Fig. 1. Represents location of sampling sites.

least to the most polluted parts.

The sampling sites are listed in table below (Table 1).

Sample collection, preparation, and data analyses

Biological data collection

Diatoms were collected as periphyton in the upper surface of selected rocks from the river with a toothbrush. Suspension materials were collected and conserved in 4% formaldehyde (Kolkwitz, 1950; Lund *et al.*, 1958; Kelly and Whitton, 1995). Samples rich in organic and inorganic matter in addition to diatoms need stronger oxidation than clean ones and the optimum ratio of sample to oxidants may need to be tested first.

Diatoms frustules were cleaned from both organic and inorganic materials by boiling in concentrated HCl (20 ml), followed by boiling in concentrated H₂SO₄ (20 ml) containing a few crystals of KNO₃, washing with distilled water following the books of Krammer and Lange-Bertalot (1986-2001). To get credible results (95%), about 500 valves were counted per each slide using 100x oil immersion, (Lund *et al.*, 1958; Pascher, 1976; Kelly and Whitton 1995; Lavoie *et al.* 2011). Diatoms were identified using the following literature: Cleve-Euler (1951-1955); Pascher (1976); Krammer and Lange-Bertalot (1986-2001); Kupe *et al.*, (2013). The trophic diatom index

(TI_{DIA}) and saprobic index (SI) for the diatoms were calculated using the formula of Shannon and Weaver (1949); Rott *et al.* (1997, 1999) as below:

$$H' = -\sum_{i=1}^n p_i \log_2(p_i) \quad .. (1)$$

$$TI_{DIA} = \frac{\sum_{i=1}^n TW_i G_i p_i}{\sum_{i=1}^n G_i p_i} \quad .. (2)$$

$$SI = \frac{\sum_{i=1}^n S_i G_i H_i}{\sum_{i=1}^n G_i H_i} \quad .. (3)$$

Where:

- H' is diversity index, with p_i denoting the proportion in group n .
- TI_{DIA} , trophic index for diatoms; TW_i , trophic value of i species (1-3); G_i , indicative weight of i species (1-3); H_i , relative frequency of i species (%); n , total number of species. Respective values of TW_i and G_i , as well as trophic classes were taken after Rott *et al.* (1999).
- SI , saprobic index for diatoms; S_i , saprobic value of i species (1-5); G_i , indicative weight of i species (1-3); H_i , relative frequency of i species (%); n , total number of species. Respective values of S_i and G_i , as well as saprobic classes were taken after Rott *et al.* (1997).

Cleaned of cell contents and mounted in a medium with a high refractive index, diatoms were identified and counted using a high power light microscope (Motic BA310, with 100X) until an appropriate sample size has been obtained. These data are then interpreted using indices or other assessment methods, according to STANDARD EN 14407 (2004). In order to rate the level of a particular stress, a biological index or score specifically designed to evaluate that stress should be used. Most of the indices were developed for use in a single geographical area but subsequent testing has shown to have broader validity. It is important that the dominant and sub-dominant taxa present in the region are represented in the index. Trophic (Rott *et al.*, 1999) and Saprobic (Rott *et al.*, 1997) indices may be used to make a rough estimate of rivers condition and to

Table 1. General sampling data from Shkumbini River (June ÷ September 2015)

Nr.	Site name	River	Code	Month	Year	Longitude	Latitude
1	Peqin	Shkumbin	AL _{SH} 0001	July/September	2015	19.76966	41.03672
2	Paper	Shkumbin	AL _{SH} 0002	July/September	2015	19.94164	41.05157
3	Labinot-Fushe	Shkumbin	AL _{SH} 0003	July/September	2015	20.1668	41.14826
4	Dardhe	Shkumbin	AL _{SH} 0004	July/September	2015	20.37328	41.1498
5	Karkavec	Shkumbin	AL _{SH} 0005	July/September	2015	20.4742	41.05906

determinate a water quality. These values are correlated to the known trophic and saprobic classes, respectively (Table 2).

Chemical data collection

For the determination of total nitrogen (TN) and total phosphorus (TP), the method of UNEP/MAP/MED POL (2005) was used. For a 50 ml of sample, 5 ml of the oxidizing agent is added. The flask and autoclave were closed for at least half an hour. Then, the autoclave was opened to allow cooling the flask nearly the room temperature. With a bulb pipette was taken from the flask 5 ml of the digested sample, on which TN as nitrate can be determined separately. The remaining 50 ml of sample was transferred to a phosphate reaction flask and added 1 ml of ascorbic acid and mixed with reagent solutions. The absorbance was measured at 880 nm after 10-30 minutes using 10 cm cell (or the longest available cell).

Results

Biological data

115 species of diatoms were found in ten sampling sites and two periods for Shkumbini River (both in July and September), (Fig. 2). Only eight of them were centric diatom, the rest were Pennate. The number of species oscillated from 28 (in Peqin, July) to 66 (in Prrenjaskarkavec, July). Other stations like: Paper and Labinot-Fushe were rich in species. Most common and widespread species of pennate diatoms were: *Cocconeis pediculus*, *Cymbella microcephala*, *Fragilaria capucina*, *Fragilaria ulna*, *Gomphonema minutum*, *Navicula cryptotenella*,

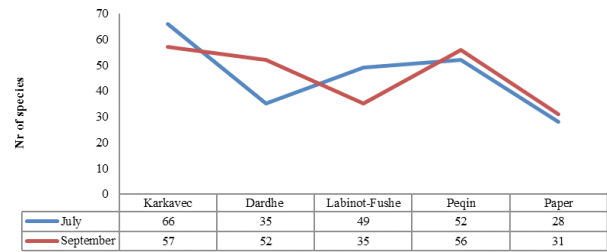


Fig. 2. Number of species at the 10 sampling sites, during sampling months (July and September 2015).

Nitzschia incospicua, *Nitzschia palea*, *Nitzschia umbonata* etc.

Diversity index (see Fig. 3) in Shkumbini river in July oscillated from 1.57 (Peqin) to 3.56 (Prrenjaskarkavec), but in September, the minimum value was found in Labinot-Fushe (1.32) and the maximum value in Dardha and Librazhd (3.74).

Regarding trophic index (TI_{DIA}) the average lowest values were observed in Paper (approximately 1.8); Dardha and Librazhd (1.9); the water quality in Dardha belongs to meso-eutrophic classes (Rott *et al.*, 1999). Two other sampling places belong to

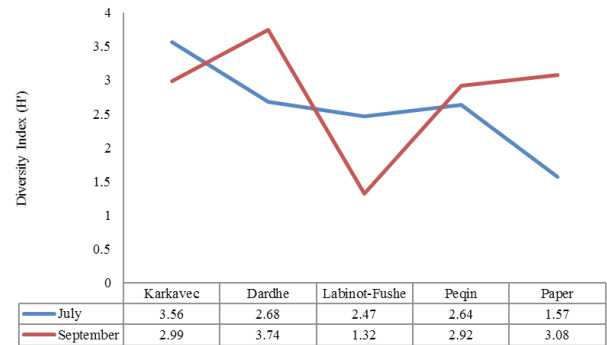


Fig. 3. Variation of Shannon Diversity Index (H'), during sampling period July-September 2015.

Table 2. Correlation between trophic and saprobic value with trophic and saprobic classes by Rott *et al.* (1997, 1999).

Value of TIDIA	Trophic classess	Description of the pollution	Value of SI	Saprobic classess	Description
£1.0	Ultraoligotroph	No polluted	1.0-<1.5	Oligosaprob	None
1.1-1.3	Oligotroph	Slightly polluted	1.5-<1.8	Oligosaprob-mesosaprob	Little
1.4-1.5	Oligo-mesotroph	Slightly polluted	1.8-<2.3	β-mesosaprob	Moderate
1.6-1.8	mesotroph	Moderately polluted	2.3-2.7	β-α mesosaprob	Fair
1.9-2.2	meso-eutroph	Obvious	2.7-<3.2	α-mesosaprob	Obvious
2.3-2.6	eutroph	Heavely polluted	3.2-<3.5	a-meso-polysprob	Heavy
2.7-3.1	eu-polytroph	Heavily to masive polluted	3.5-<4	polisaprob	Strong to
3.2-3.4	polytroph	Massive polluted			
>3.4	poly-hypertroph	Very masive polluted			

eutrophic classes (Rott *et al.*, 1999; Kupe *et al.*, 2010; Kupe *et al.*, 2013) approximately 2.6 (Karkavec and Prrenjas) and 2.5 (Labinot-Fushe). Poly trophic classes were observed for Peqini station, showing a strongly polluted part. But in September the trophic diatom index oscillated from 1.9 in Prrenjas and Karkavec to 3.2 in Peqin. The same values approximately 3.0 were found in Labinot-Fushe and in Paper. Considering the saprobic value oscillated from 1.6 (oligo β -mesosaprob) in Paper and Dardha (Librazhd) to 2.2 (β -mesosaprob to α -mesosaprob) in Peqin (Fig. 4), which show the low and the high value of saprobic index (Sládeček, 1986; Rott *et al.*, 1997). Higher densities were observed especially in Peqin and Prrenjas (Karkavec). In September, the saprobic value oscillated from 1.8 (Labinot-Fushe) to

2.0 in Peqin. In other sampling places were found 1.9 (meso-eutroph). All samples belong to the same interval in saprobic classes (β -mesosaprob) with moderate organic matter. Composition of diatom community in Shkumbini river were also classified from mesotroph (Paper) to polytroph (Peqin), by Rott *et al.* (1999) and especially Peqini site show a high level of inorganic and organic matter. Based on the trophic index in July, the water quality in Shkumbini river were classified from mesotrophic (Paper) to polytrophic (Peqin) by Rott *et al.* (1999) showing a high status of pollution.

Chemical data

The parameters which affect the water quality, nitrogen, phosphorus, and dissolved oxygen, are differ-

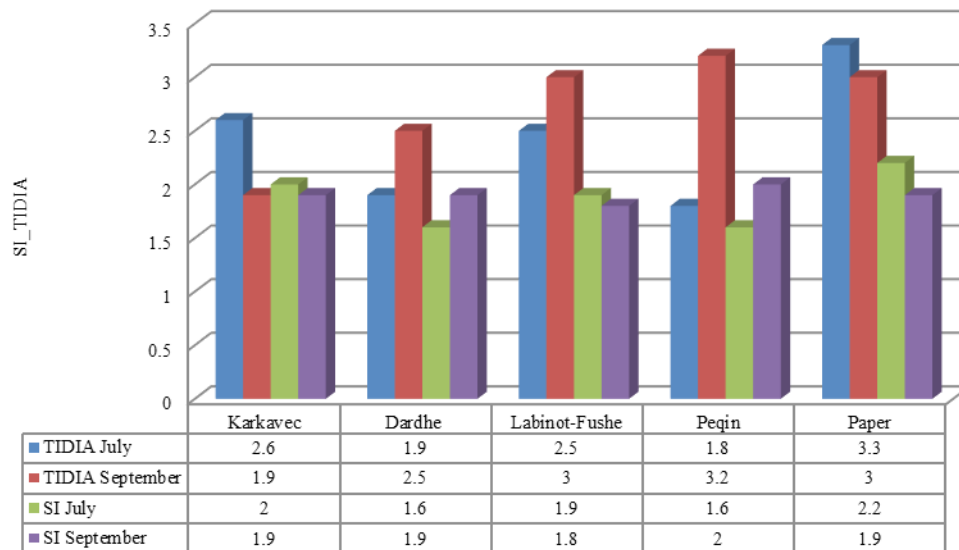


Fig. 4. Variation of Trophic diatom index (TI_{DIA}) and Saprobic index (SI) during the sampling period July-September 2015.

Table 3. The physic-chemical parameters of water and nutrients (as an average value $n=3$) in sampling stations at Shkumbini River (July-September 2015)

Sampling stations	Sampling time	DO (ppm)	ToC	pH (mS/cm)	Conductivity (mg/l)	Plot	Ntot
Karkavec	July	7.9	17.9	6.9	414	0.01c	0.19c
	September	8.8	19	7.1	423	0.012c	0.24c
Dardhe	July	7.9	21	7	398	0.016c	0.22c
	September	8.8	26.3	7.2	395	0.014c	0.34c
Labinot fushe	July	7.88	20.2	7.1	270	0.033b	0.29b
	September	9.2	27.5	6.05	262	0.05b	0.38b
Paper	July	7	18.5	6.85	340	0.044a	0.42a
	September	8.1	26.2	7	330	0.08a	0.47a
Peqin	July	7.1	18.7	6.8	475	0.055a	0.479a
	September	8	25.9	6.9	450	0.089a	0.5a

ent at stations, (Table 3). Water temperature varies according to the seasons and the sampling site from 17.9° to 21 °C in July and 19° to 27.5 °C in September. Regarding dissolved oxygen, which is an essential parameter for life of water species and organic substance presence, the conditions for living species in the stations are better in Prrenjas, Dardhe and Labinot Fushe than in more agricultural regions like Paper and Peqin. According to the trophic situation classification of fresh water by OECD, (Table 4), it results as follows: considering the content level of two main nutrients, nitrogen and phosphorous, stations Prrenjas, Dardhe, Labinot Fushe present an almost oligotrophic situation, whereas Paper and Peqin station present mesotrophic situation. *P* total, (OECD 1982; Correll 1998) for stations Peqin, Paper, Labinot field and *N* total for Prrenjas were the highest in September (the early rainy season), and lowest in July (the dry season) (Table 2).

Table 4. Classification of natural water quality (classification of fresh water by OECD)

	Oligotrophic	Mesotrophic	Eutrophic	Hypertrophic
<i>P</i> total (mg/l)	<0.013	<0.040	<0.100	>0.100
<i>N</i> total (mg/l)	<0.300	<0.400	<1.00	>1.00

Alabaster *et al.* (1980) and Abulude *et al.* (2002) considered pH range of 6.50 ÷ 9.00 as an indicator of a good fish population. In Shkumbini River pH varies from 6.05 to 7.2. The data of nutrient (*N*; *P*) concentrations in water of Shkumbini River were subjected to *t*-tests for equality of means to check the differences across the stations and seasons. Values with the different letter indicate significant differences for nutrient concentration between stations and inverse for values with the same letter, according to *t*-test for equality of means (Table 5).

Discussion

Biological interpretation

Many diatoms species in Shkumbini river were com-

mon and with not a taxonomic interest. We will be mentioning the low presence of *Caloneis* sp. that often is found in rivers in the less polluted parts as epiphyte over *Chladophora*, which were found only in Dardha (Librazhd) and Paper (in September); in other sampling places were found but not more widespread. Evaluation of the relative abundance of *Achnantheidium minutissimum* is well known (Stevenson and Bahls, 1999) and their expanses show the level of disturbance. *A. minutissima* is considered as a tolerant species (Hofmann, 1994; Miho *et al.*, 2005; Kupe, 2006) and was found almost in all sampling stations, it was often accompanied with *Cocconeis pediculus*. A special structure of diatom community was observed in the most polluted parts, like Peqini, Prrenjasi and Labinot-Fushe. *N. paleavar. palea*, as a saprotroph species (Hofmann, 1994; MIHO *et al.*, 2005), which grow up until polysaprobic habitats, was very frequent in Peqin (July) and the same species were more dominant in Dardha and Librazhd (September). The high dominance of *Nitzschia palea* together with the rare number of taxa was the causes of the low values of diversity index calculated in those stations (Fig. 3). *N. palea* was mostly associated with other saprotroph or tolerant species, like *Navicula accomoda*, *Gomphonema parvulum*, *Navicula cryptotenella*, *Fragilaria ulna*, etc. (Table 6/a and 6/b). Usually, diversity index is highly correlated with the number of species. The low value of diversity index shows the low number of common species. The most polluted part of river has indication from restaurants, agriculture land, discharge of waste and erosion by inhabitant centre of Librazhd, Prrenjas and Elbasan cities and their villages. These values show a highest pollution with organic matters. The Shannon Diversity Index was found a good and widely used tool in comparing diatoms communities' independently from pollution effects. The Saprobic Index (Rott *et al.*, 1997; Miho *et al.*, 2005; Kupe *et al.*, 2013) and Trophic Indices (Lund *et al.*, 1958; Kupe *et al.*, 2008) are correlated respectively with organic and inorganic matter in the water (Lavoie *et al.*, 2011). Furthermore, Saprobic

Table 5. Significant differences for nutrient concentration in water of Shkumbini river between seasons according to *t*-test for equality of means.

Sampling stations	Peqin			Paper			Labinot-Fushe			Dardhe			Karkavec		
	t	df	sig	t	df	sig	t	df	sig	t	df	sig	t	df	sig
Ptot (mg/l)	-5.4	4	.006*	-12.4	4	0.0001*	-4.45	4	.011*	-1.05	4	0.35	-1.93	4	0.12
Ntot (mg/l)	-0.99	4	0.37	-1.4	4	0.21	-2.36	4	0.077	-0.61	4	0.57	-3.55	4	0.02*

Table 6a. Diatom list of sampling stations in Shkumbini River (July 2015)

Sampling stations	Paper	Dardhe	Karkavec	Labinot-Fushe	Peqin	%
	p%	p%	p%	p%	p%	
Centric						
<i>Cyclotella meneghiniana</i> Kützing			0.6	0.3		40
<i>Cyclotella ocellata</i> Pantocsek	+					20
<i>Melosira varians</i> Agardh	+		0.1	0.6		60
<i>Pantocsekielladelicatula</i> (Hustedt) K.T.Kiss & E.Ács	+			+		40
<i>Thalassiosira weissflogii</i> (Grunow) Fryxell et Hasle			+	+		40
Pennate						
<i>Achnanthes minutissima</i> Kützing agg.	14.2	33.8	24.4	8.9	1.8	100
<i>Amphipleura pellucida</i> Kützing		0.2				20
<i>Amphora montana</i> Krasske			+			20
<i>Amphora ovalis</i> (Kützing) Kützing			+			20
<i>Amphora pediculus</i> (Kützing) Grunow	0.3	0.7	+			60
<i>Aneumastus stroesei</i> Oestrup	+					20
<i>Brachysira neoexilis</i> Lange-Bertalot	+			+		40
<i>Caloneis hendey</i> Lange-Bertalot			+			20
<i>Caloneis silicula</i> (Ehrenberg) Cleve			+			20
<i>Caloneis</i> sp2. (like in Brari)	+		+	+	+	80
<i>Cocconeis pediculus</i> Ehrenberg	14.7	22.2	15.4	0.3	+	100
<i>Cocconeis placentula</i> Ehrenberg agg.	1.5		3.1	1.6	+	80
<i>Cocconeis placentula</i> var. <i>lineata</i> (Ehrenberg) Van Heurck		2.7				20
<i>Cymatopleura solea</i> (Brebisson) W. Smith	+		0.3			40
<i>Cymbella affinis</i> Kützing agg.	5.6	2.4	2.3	1.3		80
<i>Cymbella amphicephala</i> Naegeli	+	+	+		+	80
<i>Cymbella caespitosa</i> (Kützing) Brun	1.0		+			40
<i>Cymbella cf. cistula</i> (Ehrenberg) Kirchner	+	+		+		60
<i>Cymbella cistula</i> (Ehrenberg) Kirchner agg.	+	+				40
<i>Cymbella descripta</i> (Hustedt) Krammer & Lange-Bertalot	+					20
<i>Cymbella helvetica</i> Kützing		0.2		+	+	60
<i>Cymbella lanceolata</i> (Ehrenberg) Van Heurck	+					20
<i>Cymbella microcephala</i> Grunow gr.	13.5	22.3	0.3	0.6	+	100
<i>Cymbella minuta</i> Hilse		+				20
<i>Cymbella silesiaca</i> Bleisch	+		0.6	1.3		60
<i>Cymbella sinuata</i> Gregory		0.3	+	+		60
<i>Cymbella subhelvetica</i> Krammer	+					20
<i>Cymbella tumida</i> (Brebissoni) Van Heurck	+		+	0.6		60
<i>Cymbella ventricosa</i> Agardh	+					20
<i>Diatoma ehrenbergii</i> Kützing	1.7		+			40
<i>Diatoma mesodon</i> (Ehrenberg) Kützing	+					20
<i>Diatoma moniliformis</i> Kützing	2.6		+	0.3	2.7	80
<i>Eunotia</i> sp.			+			20
<i>Fragilaria capucina</i> Desmazières agg.	40.1		1.5	2.5		60
<i>Fragilaria capucina</i> var. <i>perminuta</i> (Grunow) Lange-Bertalot		4.4				20
<i>Fragilaria ulna</i> (Nitzsch) Lange-Bertalot agg.	2.0	2.2	2.8	40.1	0.7	100
<i>Frustulia vulgaris</i> (Thwaites) De Toni			+			20
<i>Gomphonema angustum</i> (Kützing) Rabenhorst				+		20
<i>Gomphonema clavatum</i> Ehrenberg	+		+			40
<i>Gomphonema exilissimum</i> (Grunow) Lange-Bertalot	+					20
<i>Gomphonema gracile</i> Ehrenberg				+		20
<i>Gomphonema minutum</i> (Agardh) Agardh agg.	0.3	3.6	15.1	9.6		80
<i>Gomphonema olivaceum</i> (Hornemann) Brebisson gr.	0.7				0.4	40

Table 6a. Continued ...

Sampling stations	Paper	Dardhe	Karkavec	Labinot-Fushe	Pegun	%
	p%	p%	p%	p%	p%	
Centric						
<i>Gomphonema parvulum</i> Kützing agg.			3.3	7.3	0.4	60
<i>Gomphonema pumilum</i> (Grunow) Reichardt & Lange-Bertalot			1.2	+		40
<i>Gomphonema tergestinum</i> Fricke	0.7	1.7	+	1.0		80
<i>Gyrosigma acuminatum</i> (Kützing) Rabenhorst			+	+		40
<i>Gyrosigma scalproides</i> (Rabenhorst) Cleve	+	+	+	+		80
<i>Hantzschia amphioxys</i> (Ehrenberg) Grunow			+			20
<i>Craticula accomoda</i> Hustedt					0.4	20
<i>Craticula ambigua</i> Ehrenberg			+	+		40
<i>Craticula cuspidata</i> Kützing			+			20
<i>Fallacia pygmaea</i> Kützing			+	+	0.4	60
<i>Luticula kotschyi</i> Grunow			+			20
<i>Sellaphora bacillum</i> Ehrenberg	+					20
<i>Navicula atomus</i> (Kützing) Grunow			2.3		0.4	40
<i>Navicula capitatoradiata</i> Germain	+	0.3	7.7	1.0	0.4	100
<i>Navicula caterva</i> Hohn & Hellerman		+	+			40
<i>Navicula cryptotenella</i> Lange-Bertalot	0.3	2.0	4.4	9.9		80
<i>Navicula cryptotenelloides</i> Lange-Bertalot		+	+	+	+	80
<i>Navicula decussis</i> Oestrup	+					20
<i>Navicula gregaria</i> Donkin	0.3					20
<i>Navicula oligotraphenta</i> Lange-Bertalot & Hofmann	+	+	0.3	+		80
<i>Navicula recens</i> (Lange-Bertalot) Lange-Bertalot	+					20
<i>Navicula reichardtiana</i> Lange-Bertalot		0.3	0.6	0.3		60
<i>Navicula schroeteri</i> Meister			0.3			20
<i>Navicula schroeterii</i> Meister var. <i>simetrica</i>	+			0.6		40
<i>Navicula tripunctata</i> (O. F. Müller) Bory		+				20
<i>Navicula veneta</i> Kützing					0.4	20
<i>Navicula viridula</i> var. <i>rostellata</i> (Kützing) Cleve	+	+	0.3	+		80
<i>Neidium dubium</i> (Ehrenberg) Cleve	+		+			40
<i>Nitzschia acicularioides</i>				0.3		20
<i>Nitzschia amphibia</i> Grunow			0.9			20
<i>Nitzschia constricta</i> (Kützing) Ralfs					+	20
<i>Nitzschia denticula</i> Grunow		+	+			40
<i>Nitzschia dissipata</i> (Kützing) Grunow	+	+	1.2	1.3	1.8	100
<i>Nitzschia fonticola</i> Grunow			1.2			20
<i>Nitzschia incospicua</i> Grunow			6.1	5.7	0.7	60
<i>Nitzschia lacuum</i> Lange-Bertalot			+			20
<i>Nitzschia linearis</i> (Agarth) W. Smith var. <i>linearis</i>	+		1.2	+		60
<i>Nitzschia palea</i> (Kützing) W. Smith var. <i>palea</i>	0.5	0.7	2.6	4.1	59.8	100
<i>Nitzschia sigmoidea</i> (Nitzsch) W. Smith		+	+			40
<i>Nitzschia umbonata</i> (Ehrenberg) Lange-Bertalot					29.9	20
<i>Nitzschia valdestriata</i> Aleem et Hustedt			+		+	40
<i>Rhoicosphaenia abbreviata</i> (Agardh) Lange-Bertalot			+	+		40
<i>Rhopaloida gibba</i> (Ehrenberg) O. Müller					+	20
<i>Surirella angusta</i> Kützing	+	+	+			60
<i>Surirella bifrons</i> Ehrenberg				+		20
<i>Surirella brebissoni</i> Krammer & Lange-Bertalot			0.1	0.3		40
<i>Surirella splendida</i>			+	+		40
Number of species, N:	48	31	62	45	24	
Diversity Index (Shannon Index), H':	2.64	2.68	3.56	2.47	1.57	

Table 6a. Continued ...

Sampling stations	Paper	Dardhe	Karkavec	Labinot-Fushe	Peqin	%
	p%	p%	p%	p%	p%	
<i>Centric</i>						
Trophic Diatom Index, TI _{DIA} :	1.8	1.9	2.6	2.5	3.3	
Trophic Classes	Mesotroph	Meso-Eutroph	Eutroph	Eutroph	Polytroph	
Saprobic Index, SI	1.6	1.6	2.0	1.9	2.2	
Saprobic Classes	Oligo to β-mesosaprob	Oligo to β-mesosaprob	β-mesosaprob	β-mesosaprob	β-α-mesosaprob	

Index SI classifies water quality in rivers based on the load of dissolved organic matters originating from wastewater. The water quality assessment based on chemical-physical parameters and nutrients content are in good accordance with those obtained from biological parameters. Also, according NIVA classification and the European Directive on "Quality of fresh waters", we can conclude that water quality of station Prrenjas > Qukes > Labinot-Fushe results better than Paper and Peqin stations. It is well known that diatoms are very sensible to environmental changes like environmental conditions and pollution (Rott *et al.*, 1997, 1999).

Chemical interpretation

The recorded values are within the acceptable limits for the survival, metabolism, and physiology of aquatic organisms (Alabaster and Lloyd, 1980). This is as well reflected in nitrogen and phosphorous nutrients content that is related to the presence of organic substances which at stations Prrenjas > Dardhe > Labinot-Fushe with a higher amount of dissolved oxygen results at lower level compared to Paper and Peqin stations. The *t*-tests for mean values showed significant variation between seasons for *P* in all stations except Prrenjas and inverse for Nitrogen. *T*-tests showed that in general there was significant statistical difference between industrial and ultramafic area for the concentration of *N* and *P* in water of Shkumbini River (Table 5). It is well known that phosphorous is a necessary element for organisms' growth and usually it is the main nutrient that restricts primary productivity (phytoplankton) at a water setting. This results from the relation of this element with carbon at living organisms, through

photosynthesis process. Regarding the above-mentioned factors and the presence of dissolved oxygen in water setting of stations, it is likely suggested that the species present in stations Prrenjas > Dardhe > Labinot-Fushe, are larger in number compared to those in station Paper and Peqin. So, from biotic index parameter used for water quality, bio classification result of stations Prrenjas > Dardhe > Labinot-Fushe are better than Paper and Peqin one.

Conclusion

Depending on the total phosphorus and nitrogen in the Shkumbini River it can be classified as oligotrophic in Prrenjas, Dardhe, Labinot Fushe stations and mesotrophic in Paper and Peqin stations. The water quality of Shkumbini river in Paper and Peqin stations are suggested to affect the health and the different components of the aquatic ecosystem. The most common and widespread species of pennate diatoms were: *Achnanthes minutissima*, *Cocconeis pediculus*, *Cymbella microcephala*, *Fragilaria capucina*, *Fragilaria ulna*, *Gomphonema minutum*, *Navicula cryptotenella*, *Nitzschia incospicua*, *Nitzschia palea*, *Nitzschia umbonata* etc. *A. minutissima*, considered as a tolerant species (Cleve-Euler, 1951-1955) and was found almost in all sampling stations, it was often accompanied with *Cocconeis pediculus*. Based on diatom community structure, Shkumbini river were classified from mesotrophic (Paper) to polytrophic (in Peqin) in July, by Rott *et al.* (1999), show the high status of pollution, which indicated from medium inorganic and organic matter, originated from agricultural activity and restaurants. Especially, Peqini site show the highest level of inorganic and organic

Table 6b. Diatom list of sampling stations in Shkumbini River (September 2015)

Sampling stations	Karkavec	Labinot-Fushë	Dardhe	Peqin	Paper	%
	p%	p%	p%	p%	p%	%
Centric						
<i>Cyclotella cyclopuncta</i> Hackansson		+				20
<i>Cyclotella meneghiniana</i> Kützing		0.5		+		40
<i>Cyclotella ocellata</i> Pantocsek	3.2	+			+	60
<i>Melosira varians</i> Agardh	0.1	0.9	+	8.1	1.4	100
<i>Pantocsekielladelicatula</i> (Hustedt) K.T.Kiss & E.Ács	+	+		+		60
Pennate						
<i>Achnanthes exilis</i>	+		+			40
<i>Achnanthes minutissima</i> Kützing agg.	11.1	51.9	7.7	+	+	100
<i>Amphipleura pellucida</i> Kützing	0.4					20
<i>Amphora montana</i> Krasske			0.4			20
<i>Amphora ovalis</i> (Kützing) Kützing	+					20
<i>Amphora pediculus</i> (Kützing) Grunow	0.3		+	+		60
<i>Aneumastus stroesei</i> Oestrup	+					20
<i>Brachysira procera</i> Lange-Bertalot	5.8					20
<i>Caloneis cf. lancetula</i>			+			20
<i>Caloneis bacillum</i> (Grunow) Cleve	+					20
<i>Caloneis</i> sp.		+	0.7	+	3.2	80
<i>Cocconeis pediculus</i> Ehrenberg	2.3	+	+	+	+	100
<i>Cocconeis placentula</i> Ehrenberg agg.	+	2.4	+	13.5	11.8	100
<i>Cymatopleura solea</i> (Brebisson) W. Smith			+	+		40
<i>Cymbella affinis</i> Kützing agg.	0.5		7.4	+		60
<i>Cymbella amphicephala</i> Naegeli	+		+	+		60
<i>Cymbella caespitosa</i> (Kützing) Brun	1.1		+	0.9		60
<i>Cymbella cistula</i> (Ehrenberg) Kirchner agg.	+		+	0.4	0.4	80
<i>Cymbella descripta</i> (Hustedt) Krammer & Lange-Bertalot	0.3					20
<i>Cymbella helvetica</i> Kützing	0.3					20
<i>Cymbella lanceolata</i> (Ehrenberg) Van Heurck			+	+		40
<i>Cymbella microcephala</i> Grunow gr.	8.5	+	3.2	+	+	100
<i>Cymbella minuta</i> Hilse					+	20
<i>Cymbella silesiaca</i> Bleisch	+		1.1	+	0.7	80
<i>Cymbella sinuata</i> Gregory			+	+		40
<i>Cymbella subhelvetica</i> Krammer	+		+			40
<i>Cymbella tumida</i> (Brebissoni) Van Heurck		+		1.1	1.8	60
<i>Cymbella turgidula</i> (cf. var. <i>venezuelana</i> Krammer)			1.4			20
<i>Denticula tenuis</i> Kützing	0.5					20
<i>Diatoma ehrenbergii</i> Kützing	3.7		+			40
<i>Diatoma mesodon</i> (Ehrenberg) Kützing			+			20
<i>Diatoma moniliformis</i> Kützing	36.7	+	3.2	0.7	0.7	100
<i>Diatoma vulgare</i> Bory gr.			+	0.4		40
<i>Fragilaria acus</i> Kützing		+				20
<i>Fragilaria arcus</i> (Ehrenberg) Cleve	+					20
<i>Fragilaria biceps</i> (Kützing) Hustedt	1.1					20
<i>Fragilaria capucina</i> Desmazières agg.	17.9	2.8		+		60
<i>Fragilaria capucina</i> var. <i>rumpens</i> (Kützing) Lange-Bertalot			6.0			20
<i>Fragilaria parasitica</i> (W. Smith) Grunow var. <i>parasitica</i>	+					20
<i>Fragilaria ulna</i> (Nitzsch) Lange-Bertalot agg.	+	0.5	8.1	36.5	17.5	100
<i>Frustulia vulgaris</i> (Thwaites) De Toni			+			20
<i>Gomphonema angustum</i> (Kützing) Rabenhorst	0.3				+	40
<i>Gomphonema gracile</i> Ehrenberg	0.3	+				40

Table 6b. Continued ...

Sampling stations	Karkavec	Labinot-Fushe	Dardhe	Peqin	Paper	%
	p%	p%	p%	p%	p%	
Centric						
<i>Gomphonema minutum</i> (Agardh) Agardh agg.	1.1	6.1		2.2		60
<i>Gomphonema olivaceum</i> (Hornemann) Brebisson gr.	0.3		2.5	+	0.4	80
<i>Gomphonema parvulum</i> Kützing agg.		3.8	1.1	2.8	0.9	80
<i>Gomphonema pumilum</i> (Grunow) Reichardt & Lange-Bertalot		+	0.4	+		60
<i>Gomphonema tergestinum</i> Fricke	0.3		1.4		+	60
<i>Gomphonema truncatum</i> Ehrenberg	+			+		40
<i>Gyrosigma acuminatum</i> (Kützing) Rabenhorst	+	+		+		60
<i>Gyrosigma scalproides</i> (Rabenhorst) Cleve			0.4	1.1	1.8	60
<i>Craticula accomoda</i> Hustedt		2.8			+	40
<i>Craticula ambigua</i> Ehrenberg			+			20
<i>Craticula cuspidata</i> Kützing				+	0.4	40
<i>Sellaphora bacillum</i> Ehrenberg	+					20
<i>Navicula atomus</i> (Kützing) Grunow		2.4	0.4		1.8	60
<i>Navicula capitatoradiata</i> Germain	+		0.7	13.0	7.9	80
<i>Navicula caterva</i> Hohn & Helleman	+	+	+			60
<i>Navicula cryptocephala</i> Kützing		+				20
<i>Navicula cryptotenella</i> Lange-Bertalot	0.3	18.9	1.1	9.1	20.2	100
<i>Navicula cryptotenelloides</i> Lange-Bertalot			+	+	+	60
<i>Navicula oligotrappenta</i> Lange-Bertalot & Hofmann	+		+	+		60
<i>Navicula radiosa</i> Kützing	+					20
<i>Navicula reichardtiana</i> Lange-Bertalot	0.5		8.1	+	+	80
<i>Navicula saprophila</i> Lange-Bertalot			5.6			20
<i>Navicula schroeteri</i> Meister			0.7	+		40
<i>Navicula schroeterii</i> Meister var. <i>simetrica</i>		+		+		40
<i>Navicula tripunctata</i> (O. F. Müller) Bory	+		+	+		60
<i>Navicula veneta</i> Kützing		0.9		+		40
<i>Navicula viridula</i> var. <i>grunowii</i>			+			20
<i>Nitzschia acicularis</i> W. Smith				+		20
<i>Nitzschia amphibia</i> Grunow	+			+		40
<i>Nitzschia angustata</i> (W. Smith.) Grunow	0.3					20
<i>Nitzschia constricta</i> (Kützing) Ralfs				+		20
<i>Nitzschia denticula</i> Grunow	0.4				+	40
<i>Nitzschia dissipata</i> (Kützing) Grunow	+	+	17.3	6.7	24.6	100
<i>Nitzschia fonticola</i> Grunow	+					20
<i>Nitzschia hungarica</i> Grunow				+		20
<i>Nitzschia incospicua</i> Grunow		+		+		40
<i>Nitzschia lacuum</i> Lange-Bertalot	1.6					20
<i>Nitzschia linearis</i> (Agardh) W. Smith var. <i>linearis</i>	1.1			+		40
<i>Nitzschia palea</i> (Kützing) W. Smith var. <i>palea</i>		6.1	19.7	2.4	4.6	80
<i>Nitzschia sinuata</i> var. <i>tabellaria</i> (Grunow) Grunow	+					20
<i>Nitzschia umbonata</i> (Ehrenberg) Lange-Bertalot		+				20
<i>Nitzschia vermicularis</i> (Kützing) Hantzsch				+		20
<i>Rhoicosphaenia abbreviata</i> (Agardh) Lange-Bertalot				+		20
<i>Surirella angusta</i> Kützing			1.1	0.7		40
<i>Surirella bifrons</i> Ehrenberg				+		20
<i>Surirella brebissoni</i> Krammer & Lange-Bertalot			0.7	0.4	0.4	60
Number of species, N	53	32	48	53	29	99
Diversity Index, H'	2.99	1.32	3.74	2.92	3.08	
Trophic Diatom Index, TI _{DIA}	1.9	3.0	2.5	3.2	3.0	

Table 6b. *Continued ...*

Sampling stations	Karkavec	Labinot-Fushe	Dardhe	Peqin	Paper	
Centric	p%	p%	p%	p%	p%	%
Trophic Classes:	Meso-eutroph	Eu-Polytroph	Eutroph	Polytroph	Eu-polytroph	
Saprobic Index, SI	1.9	1.8	1.9	2.0	1.9	
Saprobic Classes	β -mesosaprob	β -mesosaprob	β -mesosaprob	β -mesosaprob	β -mesosaprob	

matter. In September is oscillated from meso-eutroph (Prrenjas-Karkavec) to polytrophic in Peqin. In this period the water of Shkumbini river was charged with moderately to heavily polluted inorganic matter and moderate polluted with organic matter. We observed, by the levels of nitrogen and phosphorus in the water, increased turbidity due to the large content of solids, in some cases, large and organic pollution, e.g., stations: the saprobic index belong to $\hat{\alpha}$ -meso-saprob (2.2) Peqin; $\hat{\alpha}$ -meso-saprob (2.0) Prrenjas and $\hat{\alpha}$ -mesosaprob (1.9) Labinot-Fushe. Presence of macroscopic algae like as *Chladophora glomerata* was not widespread (only in July). This situation comes as a result of the influence of cities: Prrenjas, Librazhd, Elbasan, Peqin and Rrogzhinë; that are the main reason of high values for trophic and saprobic index in Shkumbini river. Most of farmlands in the watershed are irrigated, which increases the load to the river. Most of the drained water discharged directly or indirectly into the river. Industry, restaurants, or other discharge also contributes pollution to the river in many forms. In Shkumbini River, the metal parts factories discharge wastes to the river without treatment. Shkumbini River is fertilized by nutrients in detergents and human and animal waste and by nutrients in runoff from the land. As a result, the river is becoming more eutrophic and polytrophic. Evidence of the ecological impacts of human activities is apparent in both the aquatic plant community and the benthic diatoms in the water.

Acknowledgement

The Critical Ecosystem Partnership Fund is acknowledged for funding of the project Integrated Natural Water Management of Shkumbini River, Albania.

References

Abulude, F.O. 2005. Nutritional evaluation of aquatic weeds in Nigeria. *J. Env. Agric. Food Chem.* 4 (1): 835-840.

Alabaster, J. S. and Lloyd, R. 1980. *Water Quality Criteria for Freshwater Fish.* London Boston.

Arle, J., Mohaupt, V. and Kirst, I. 2016. Monitoring of Surface Waters in Germany under the Water Framework Directive- A Review of Approaches, Methods and Results. *Water.* 8 : 217. <https://doi.org/10.3390/w8060217>

Carpenter, S.R., Caraco, N.F., Correll, D.L., Howarth, R.W., Sharpley, A.N. and Smith, V.H. 1998. Nonpoint pollution of surface waters with phosphorus and nitrogen. *Ecological Applications.* 8 : 559-568.

Chapman, Deborah, V., World Health Organization, UNESCO and United Nations Environment Programme. 1996. *Water quality assessments: a guide to the use of biota, sediments and water in environmental monitoring.*, E & FN Spon 2nd ed. London.

Cleve-Euler, A. 1951 – 1955. *Die Diatomeen von Schweden und Finnland, Teil (I – V).* Almqvist & Wiksells Boktryckeri AB, Stockholm.

Correll, D.L. 1998. The role of phosphorus in the eutrophication of receiving waters: a review. *Journal of Environmental Quality.* 27 : 261-266.

- Cox, E.J. 1991. What is the basis for using diatoms as monitors of river quality? *Proceedings of International Symposium, Landsat fur Wasser und Abfall Nordrhein*. 33 – 40, Dusseldorf.
- Hofmann, G. 1994. Aufwuchs-Diatomeen in Seen und Eignung als Indikatoren der trophie. - *Schweizerbart Science Publishers*, Stuttgart, Germany, 241 pp. ISBN: 9783443570217.
- Jakovljević, O.S., Popović, S.S., Vidaković, D.P., Stojanović, K.Z. and Krizmanić, J. • 2016. The application of benthic diatoms in water quality assessment (Mlava River, Serbia). *Acta Botanica Croatica*. 75 (2): 199-205.
- John, J. 2002. Bio assessment of health of aquatic systems by the use of diatom. In: MODERN TRENDS IN APPLIED ECOLOGY. Ed. By: Ambasht R.S. & Ambasht N.K., Kluwer Ac. Plenum Publishers, 1-20.
- Grasshoff, K., Ehrhardt, M. and Kremling, K. 1999. *Methods of Seawater Analysis* (3rd extended edition). John Wiley & Sons, Ltd, Weinheim.
- Guignard, M.S., Leitch, A.R., Acquisti, C., Eizaguirre, C., Elser, J.J., Hessen, D.O., Jeyasingh, P.D., Neiman, M., Richardson, A.E., Soltis, P.S., Soltis, D.E., Stevens, C.J., Trimmer, M., Weider, L.J., Woodward, G. and Leitch, I.J. 2017. Impacts of Nitrogen and Phosphorus: From Genomes to Natural Ecosystems and Agriculture. *Front. Ecol. Evol.* 5 : 70. doi: 10.3389/fevo.2017.00070
- Kabo, M. 1990-1991. *Gjeografia fizike e Shqipërisë*. Vol. I & II. Akademia e Shkencave, Tiranë.
- Kelly, M. G., Cazaubon, A., Coring, E., Dell'Uomo, A. and Ector L. 1998. Recommendations for the routine sampling of diatoms for water quality assessment in Europe. *J. Appl. Phycol.* 10 : 215-224.
- Kelly, M.G. and Whitton, B.A. 1995. The trophic diatom index: a new index for monitoring eutrophication in the rivers. *Journal of Applied Phycology*. 7 : 433-444.
- Kolkwitz, R. 1950. Ökologie der Saprobien. Über die Beziehungen der Wasserorganismen zur Umwelt. Schriftenreihe des Vereins für Wasser. *Boden und Lufthygiene*. 4 : 1-64.
- Kovacs, C., Kahlert, M. and Padisak, J. 2006. Benthic diatom communities along pH and TP gradients in Hungarian and Swedish streams. *J. Appl. Phycology*. 18 : 105-117.
- Krammer, K. and Lange-Bertalot, H. 1986-2001. *Süßwasserflora von Mitteleuropa*. Bd. 02/5: Bacillariophyceae, Stuttgart, 2/1: 876; 2/2: 596; 2/3: 576; 2/4: 437; 2/5.
- Kupe L. 2006. *Vlerësimi i gjendjes mjedisore të disa habitateve ujore shqiptare mbështetur tek diatometë*, PhD Thesis. Universiteti Bujqësor, Tirane, Albania.
- Kupe, L., Schanz, F. and Bachofen, R. 2008. Biodiversity in the Benthic Diatom Community in the Upper River Töss Reflected in Water Quality Indices. *Clean - Soil Air Water*. 36(1) : 84-91.
- Kupe, L., Poçi, A., Miho, A. and Hübener T. 2010. Microscopic algae from karst lakes of Dumre region (Central Albania). *Botanica Serbica*. 34 (2) : 87-98.
- Kupe, L., Imeri, A., Cara, M. and Kurti, D. 2013. Use of diatom and macrophyte index to evaluate the water quality in Ohrid Lake. *Journal of the Faculty of Engineering and Architecture of Gazi University*. 28(2) : 393-400.
- Lavoie, I., Campeau, S., Grenier, M. and Peter, J. D. 2011. A diatom-based index for the biological assessment of eastern Canadian rivers: an application of correspondence analysis (CA). *Canadian Journal of Fisheries and Aquatic Sciences*. 63(8) : 1793-1811.
- Lund, J. W. G., Kipling, C. and Lecren, E. D. 1958. The inverted microscope method of estimating algal numbers and the statistical basis of estimations by counting. *Hydrobiologia*. 2 : 143-170.
- Miho, A., Cullaj, A., Hasko, A., Lazo, P., Kupe, L., Schanz F., Brandl, H., Bachofen, R. and Baraj, B. 2005. *Gjendja mjedisore e disa lumenjve të Ultësirës Adriatike Shqiptare/ Environmental state of some rivers of Albanian Adriatic Lowland*. SCOPES program (Swiss National Science Foundation - SNSF), Tirana (In Albanian with a summary in English).
- OECD (Organization for Economic Cooperation and Development), 1982. *Eutrophication of waters. Monitoring, assessment and control*. Paris, France.
- Paparisto, A., Lazo, P., Halimi, E., Duka, S., Hamzaraj, E., Laknori, O. and Pepa, B. 2010. Assessment of water quality of Shkumbini river, Albania. *Asian Journal of Chemistry*. 22 : 6164-6172.
- Pascher A. 1976. *Süßwasserflora von Mitteleuropa*, Heft 10, Jena.
- Prygiel J., Carpentier., Almeida S. and Coste, M. 2002. Determination of the biological Diatom Index (IBD NFT 90-354): results of an intercomparison exercise. *J. Appl. Phycology*. 14 : 27-39.
- Rott, E., Hofmann, G., Pall, K, Pfister, P. and Pipp E. 1997. *Indikationslisten für Aufwuchsalgen in Fließgewässern in Österreich*. Teil 1: Saprobien Indication. Projekt des Bundesministeriums für Land- und Forstwirtschaft, Wasserwirtschaftskataster, 80 pp.
- Rott, E., Pipp, E., Pfister, P., Van Dam, H., Ortler, K., Binder, N. and Pall K. 1999. *Indikationslisten für Aufwuchsalgen in Österreichischen Fließgewässern*. Teil 2: Trophieindication. - Bundesministerium f. Land und Forstwirtschaft, Zahl 41.034/08-IVA 1/97, Wien, 248 pp.
- Shannon, C. E. and Weaver, W. 1949. *The mathematical theory of communication*. University of Illinois Press, Urbana.
- Sládeček V. 1986. Diatoms as indicators of organic pollution. *Acta Hydrochimica et Hydrobiologica*. 14 : 555-566.
- STANDAREN 139462003. *Water quality - Guidance standard for the routine sampling and pretreatment of benthic diatoms from rivers*, 13 pp.

- STANDARD EN 14407 2004. *Water quality*. Guidance standard for the identification, enumeration and interpretation of benthic diatom samples from running waters, 12 pp.
- Stevenson, R. J. and Bahls, L. L. 1999. Periphyton protocols. In: Barbour, M. T., Gerritsen, J., Snyder, B. D. and Stribling, J. B. 1999. *Rapid bio assessment protocols for use in streams and wade able rivers: periphyton, benthic macroinvertebrates and fish*. 2nd ed. EPA 841-B-99-002, US Environmental Protection Agency, Office of water, Washington, DC, pp. 6.1-6.22.
- UNEP/MAP/MED POL 2005. *Sampling and Analysis Techniques for the Eutrophication Monitoring Strategy of MED POL*. MAP Technical Reports Series No. 163. UNEP/MAP, Athens.
- Weigelhofer, G., Hein, T. and Bondar-Kunze, E. 2018. Phosphorus and Nitrogen Dynamics in Riverine Systems: Human Impacts and Management Options. In: Schmutz S., Sendzimir J. (eds) *Riverine Ecosystem Management*. Aquatic Ecology Series, vol 8. Springer, Cham. https://doi.org/10.1007/978-3-319-73250-3_10
- Whitton, B.A. and Rott, E. 1996. Use of Algae for Monitoring Rivers. *Proc. Internat. Symp.* Innsburck, Austria 17-19 September 1995. Instit. Für Botanik, Univ. Innsburck.
- Zelinka, M. and Marvan, P. 1961. Zur Praeisierung der biologischen Klassifikation der Reinheit fliesse. *Arch. Hydrobiology*. 57 : 389-407.